

Mapping Invasive Plant Species in the Sacramento- San Joaquin Delta Region Using Hyperspectral Imagery

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INTRODUCTION

The spread of invasive plant species poses a significant threat to biological diversity and ecosystem functioning (Mooney and Cleland 2001, Pimm and Gilpin 1989). Considering the threats to biological diversity, invasive plants have longer lasting effects than other anthropogenic activities (Bossard et al. 2000). Even after invasive plants are no longer introduced, the original populations can continue to spread and invade new areas. Compared to terrestrial weeds, invasive aquatic vascular plants are more difficult to control and manage because they grow submerged in water, float on the water surface, or have inundated basal portions with emergent foliage and upper canopy. Of particular concern in the Sacramento-San Joaquin Delta Region are Brazilian waterweed (*Egeria densa*) and water hyacinth (*Eichhornia crassipes*), which are well known for their ability to alter physical and biological functions of aquatic systems (Bossard et al. 2000).

A fundamental need for invasive aquatic plant management is to develop a cost-effective, large scale monitoring method (Williams and Hunt 2002, Johnson 1999). As a non-intrusive and repeatable mapping method, both aerial and satellite imagery have been applied to mapping invasive vegetation in various aquatic environments (Ackleson and Klemas 1987, Lehmann et al. 1997, Lehmann and Lachavanne 1997, Porter et al. 1997, Williams and Lyon 1997, Zhang 1998, Valta-Hulkkone et al. 2003, Sawaya et al. 2003, Ward et al. 2003). However, these types of data typically provide limited ability to discriminate between aquatic species due to either low spectral resolution, low spatial resolution, or both. In recent years hyperspectral remote sensing has shown promising results for species level mapping which has opened a new avenue for map invasive aquatic plants (Penuelas et al. 1993, Zhang et al. 1997, Spanglet et al. 1998, Silvestri et al. 2002, Schmidt and Skidmore 2003). The use of remotely sensed data to map invasive aquatic and upland weeds in the Sacramento-San Joaquin Delta Region was tested in a pilot project in the summer and fall of 2003 (Mulitsch and Ustin, 2003). This project successfully provided accurate, spatially explicit information about the distribution and infestation levels of invasive weeds within the test area of the Delta Region.

The results presented here build upon the methodology developed in the pilot project and extends it to the entire Sacramento-San Joaquin Delta Region, to provide the California Department of Boating and Waterways with a comprehensive baseline for documenting the distribution and extent of invasive species over the entire Sacramento-San Joaquin Delta Region as well as a rapid and valuable technique for making informed decisions relevant to managing the extent and spread of invasive species. The goals of this study were to: 1) acquire and process airborne remotely sensed hyperspectral imagery for the Sacramento-San Joaquin Delta Region; 2) conduct field surveys in support of the overflights; and 3) map the distribution and abundance of selected invasive species within the Sacramento-San Joaquin Delta Region.

Description of Target Invasive Species

In the Sacramento-San Joaquin Delta Region there are a number of highly invasive plant species which have become particularly problematic in recent years. Two species in particular are of concern to the Department of Boating and Waterways (CDBW) and the California Department of Food and Agriculture (CDFA): Brazilian waterweed and water hyacinth.

Brazilian waterweed (*Egeria densa*)

Brazilian waterweed forms thick mats of long, intertwining, multi-branched stems below the water surface. It significantly slows water flow in channels, which interferes with irrigation projects, hydroelectric utilities, and urban water supplies. Dense mats of underwater Brazilian waterweed can interfere with recreational and commercial activities within the Delta, such as boating, swimming and fishing (Bossard et al. 2000).

Water hyacinth (*Eichhornia crassipes*)

Dense areas of water hyacinth can quickly dominate waterways because of rapid leaf production, fragmentation of daughter plants, and copious seed production and germination (Bossard et al. 2000, Penfound and Earle 1948). Water hyacinth degrades habitat for waterfowl by reducing open water area used for resting. When it decomposes, the water becomes unsuitable for drinking, which further impacts the wildlife in the Delta. As with other invasive species, it quickly displaces the native aquatic plants used for food or shelter by wildlife species. The effects of this species on humans includes the obstruction of navigable waterways, impeding drainage, fouling hydroelectric generators and water pumps, and blocking irrigation channels (Bossard et al. 2000).

METHODS

Study Area Description

The Sacramento-San Joaquin Delta spans approximately 3400 km² (~840,000 acres) from Yolo County to Fresno County in California's Central Valley. Water bodies within the proposed project boundary include the Delta estuary and the Merced, San Joaquin, Sacramento, and Tuolumne Rivers. Because of the Delta's proximity to San Francisco Bay, there is a strong marine influence in the estuary creating a unique and diverse ecosystem. For the purposes of our analyses we divided the study site into two areas consisting of a northern area, referred to as Priority 1 (2,139 km², Figure 1), and Priority 2 (1,252 km², Figure 2) in the southern section.

Through the acquisition of airborne hyperspectral imagery and the collection of field data the objectives of this study were to; (1) map the two target species across the spatial extent of the Delta, (2) within a selected number of priority sites identified by CDBW (Figure 3) map and calculate the area of infestation of the two target species, and (3) investigate the potential for discriminating Brazilian waterweed from other submerged aquatic vegetation (SAV) using hyperspectral image analysis techniques.

The following section provides information on the field data collection and on the acquisition of the hyperspectral imagery. We then provide details on a number of the processing steps that are necessary to prepare the imagery before classification, such as improving the spatial accuracy of the images and the creation of masks to limit our extent of analysis within the study site, followed by a description of the techniques used to identify and map the target species. All image processing and mapping work was conducted using ENVI software (version 4, Research Systems Inc., Boulder, Colorado) and ArcMap (version 9, ESRI, Redland, California). The processing and results presented in this report refer to the Priority 1 study area.

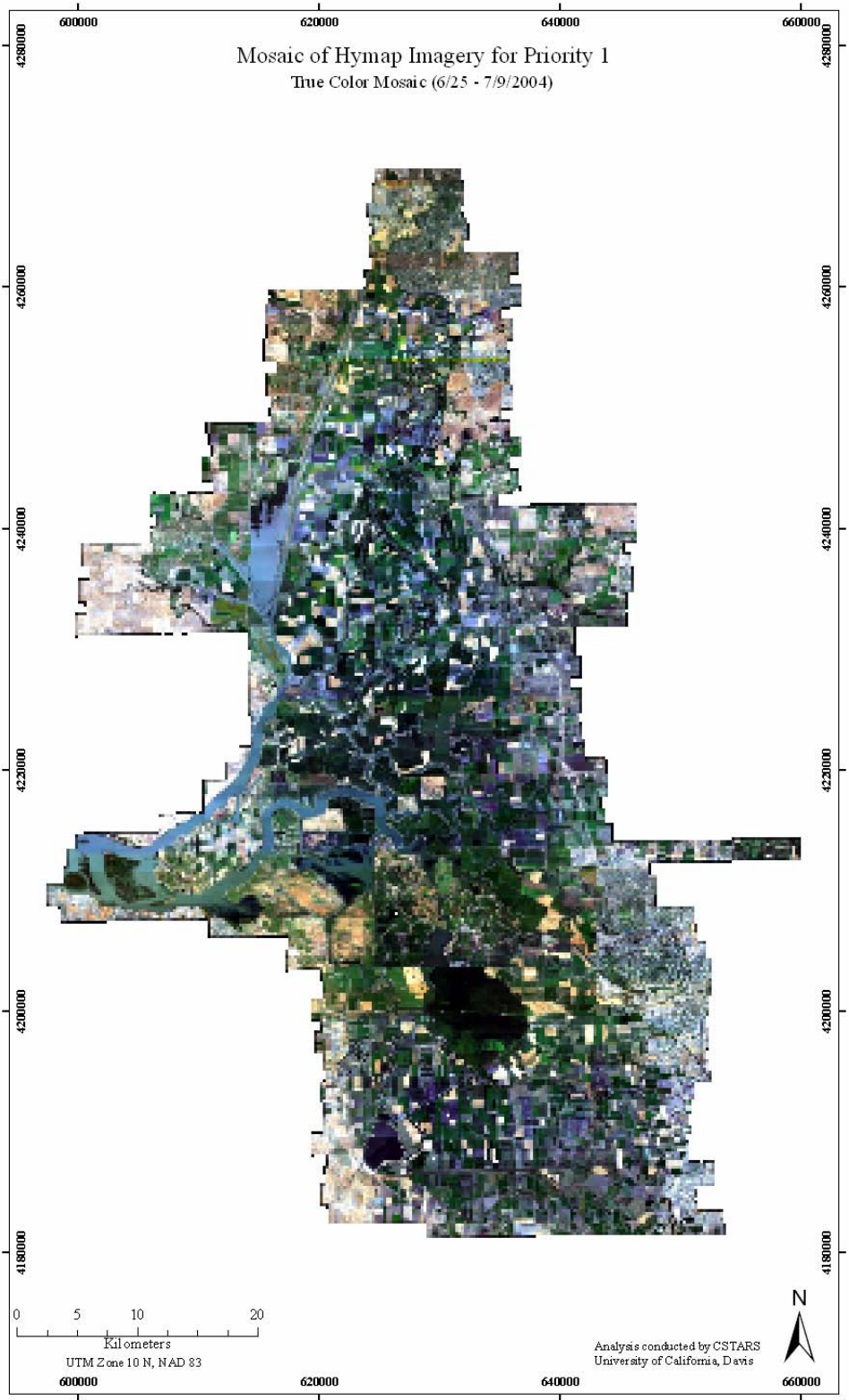


Figure1

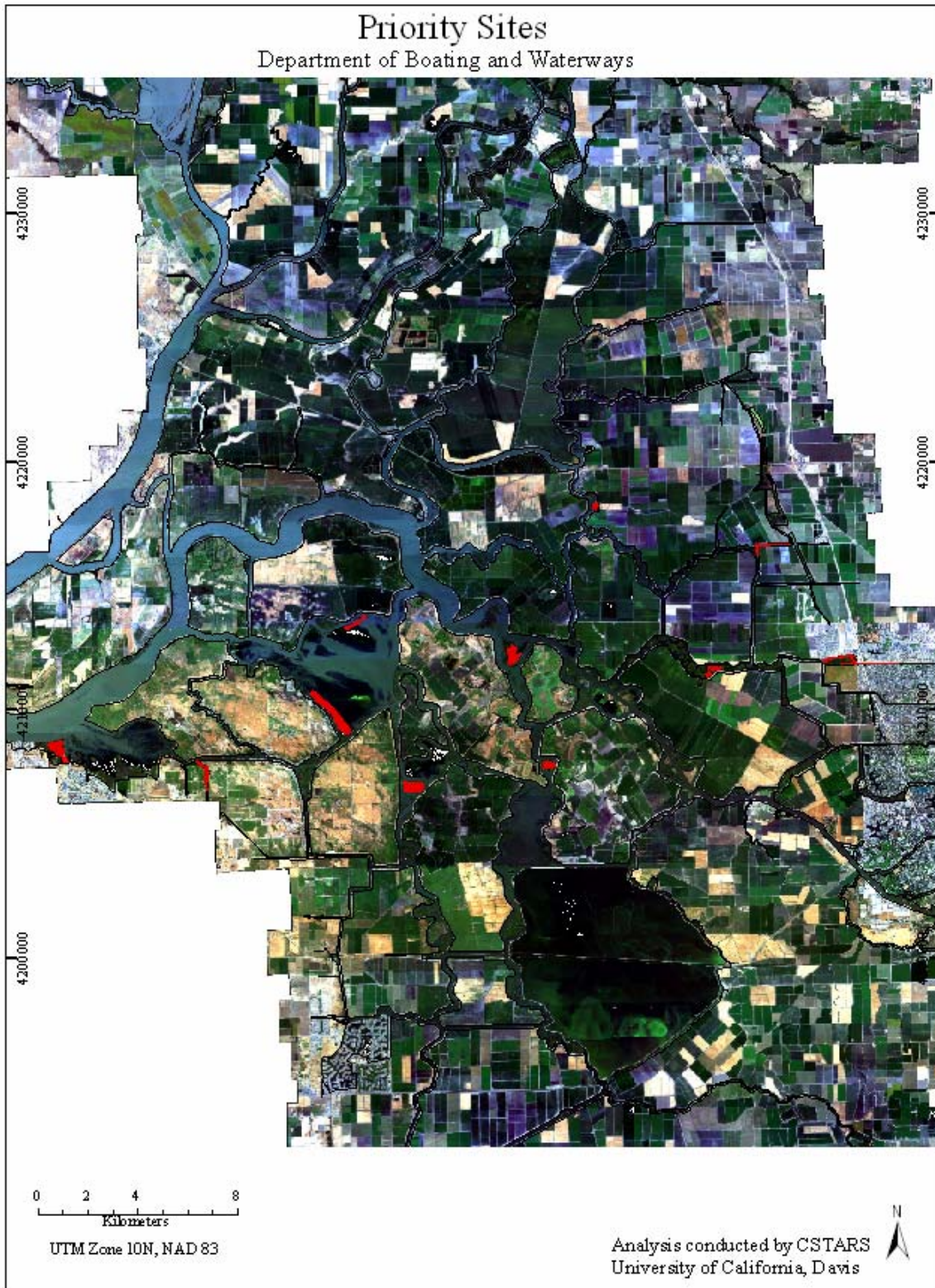


Figure 3

Fieldwork Description

Researchers from the Center for Spatial Technology and Remote Sensing (CSTARS) and California Space Institute (CalSpace) Center of Excellence, together with boatmen and botanists from CDFG conducted fieldwork between 24 June and 2 July 2004 to coincide with image acquisition. The field crew was divided into three teams that launched at Brannan Island, Paradise Point and Discovery Bay. Each team was composed of two researchers per airboat and motor boat, who collected information on six target species at both the CDBW priority sites and the surrounding river channels (Figure 4). The six primary species were; Brazilian waterweed, water hyacinth, pennywort (*Hydrocotyle ranunculoides*), water primrose (*Ludwigia peploides*), tule (*Scirpus acutus*), and cattail (*Typha latifolia*). A second tier of species included other submerged aquatic species, riparian species, and invasive levee species such as pepperweed (*Lepidium latifolium*) and purple loosestrife (*Lythrum salicaria*) where present (Appendix, Table 1). The following information was acquired when patches of the target weeds were encountered:

- Global Positioning System (GPS) reading of the patch location, the percent cover of the target species, percent cover of co-occurring species, approximate dimensions of the patch, and presence of flowers (for Brazilian waterweed or water hyacinth);
- Digital photos of the weed patch (vertically down and oblique) linked to the GPS points via the date and time stamp;
- GPS points taken of uninfested water and a Secchi depth reading; and
- One plant sample of each of the target species each day to ensure correct identification.

The majority of GPS readings were taken from large (> 1 pixel or 3 m x 3 m) homogenous patches (>75% cover) of the target species, with a minimum separation distance of 20 m. In many cases, for example, there were difficulties in getting the motorboat into a patch of weeds, so GPS points were recorded by entering distance and direction information from the actual GPS position recorded on the boat.

Image Acquisition

Airborne remotely sensed hyperspectral imagery for the Sacramento-San Joaquin Delta Region was acquired using the HyMap airborne sensor (Cocks, 1998) between 25 June and 7 July 2004 (Figure 5). The dates of image acquisition were selected to precede the control efforts by the CDBW and to coincide with peak flowering of Brazilian waterweed.

HyMap collects data over 126 spectral bands ranging from 0.45 to 2.5 μm (bandwidth 15-20 nm). The spectral range is collected using four detectors in the visible (VIS, 0.42–0.88 μm), near-infrared (NIR, 0.881–1.335 μm), shortwave-infrared (SWIR1, 1.40–1.81 μm and SWIR2, 1.95–2.49 μm). HyMap has an instantaneous field-of-view of 2.0 mrad (across track) and 2.5 mrad (along track) and was flown at an altitude of 1,500 m to produce pixels \sim 3 m on a side. The field-of-view of the sensor is 512 pixels, producing a swath of approximately 1.5 km across track (<http://www.hyvista.com>). In total, 139 images (or flightlines) were collected, 65 in the Priority 1 region (Figure 1) and

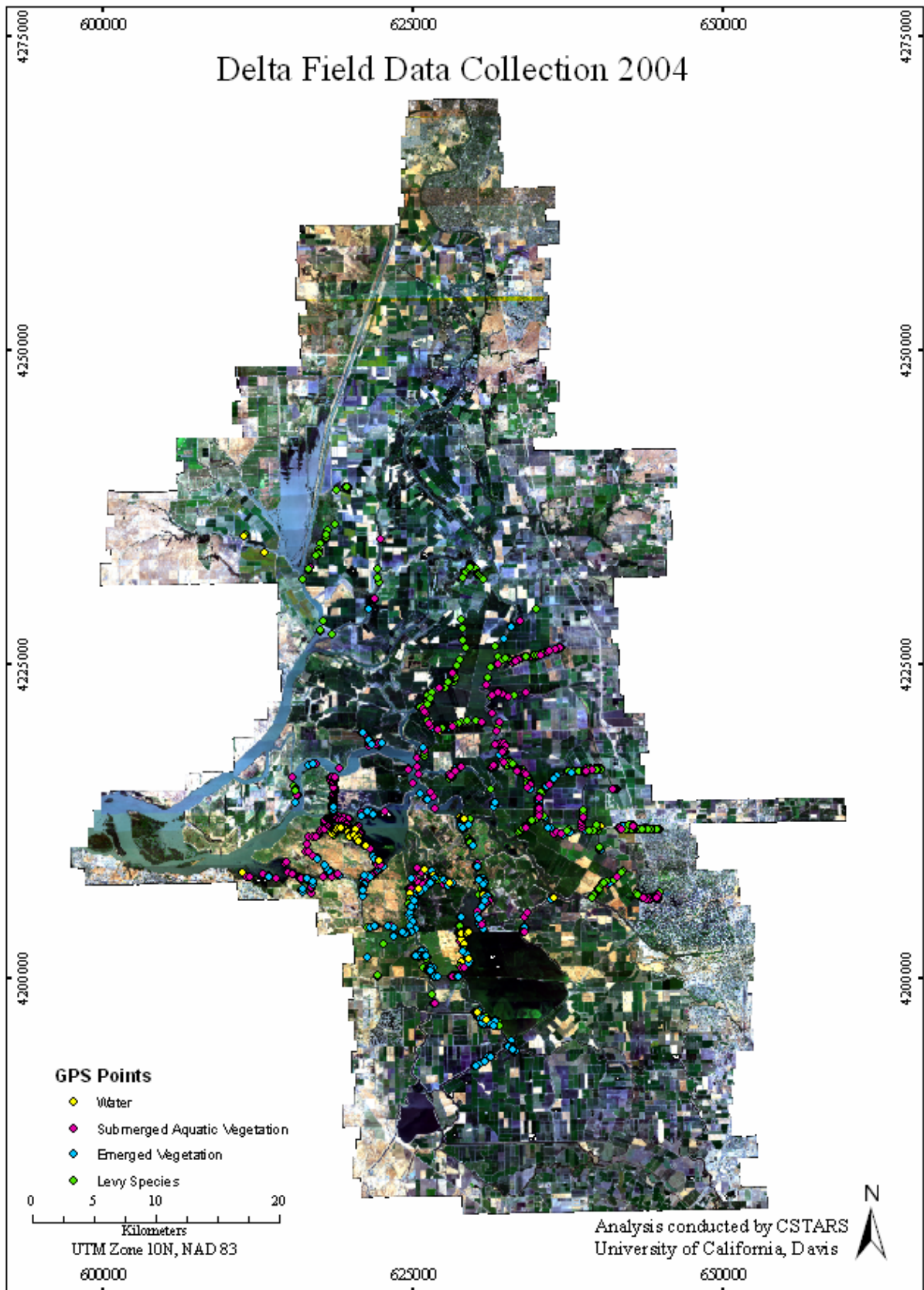


Figure 4

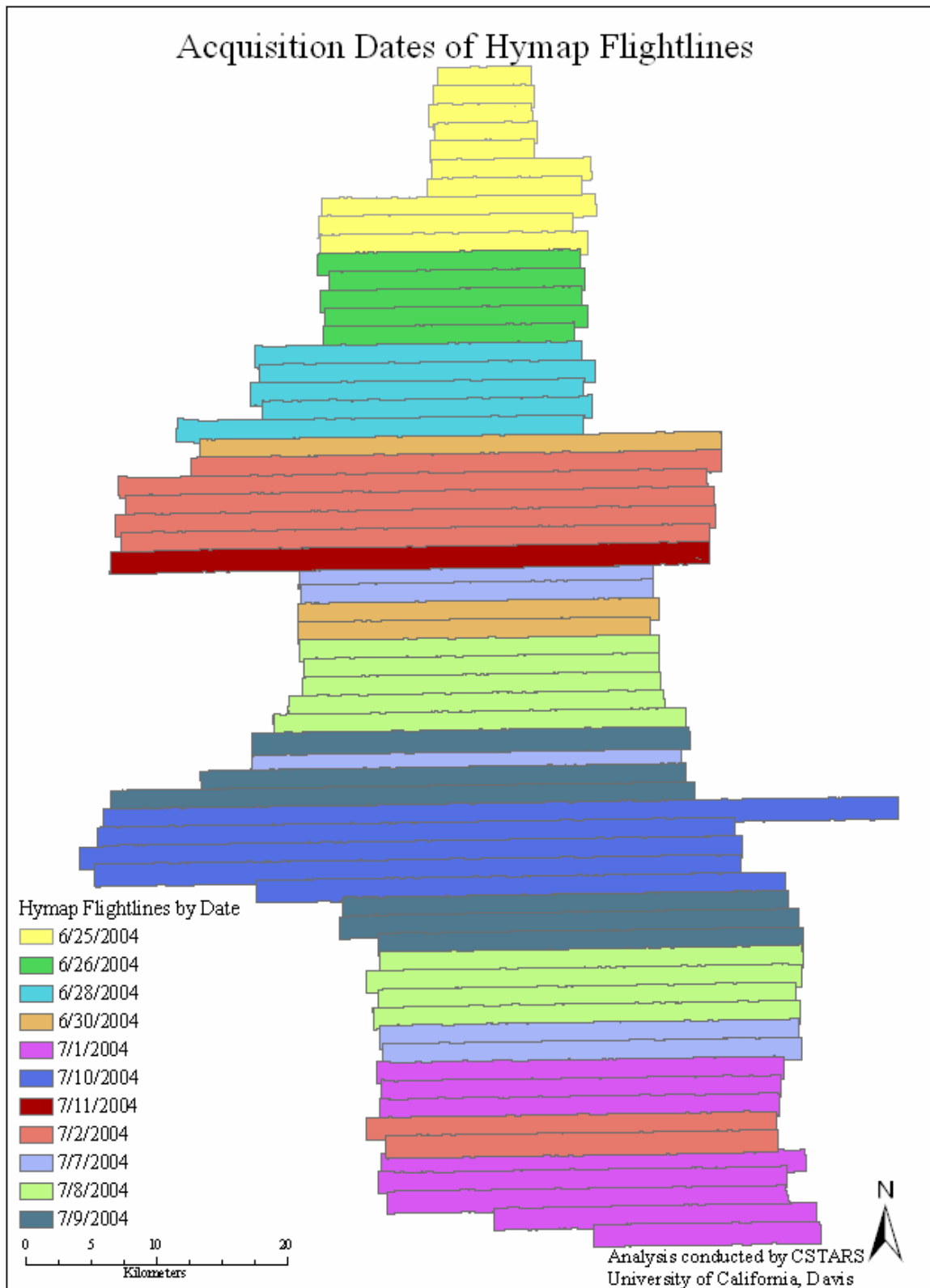


Figure 5

74 in the Priority 2 region (Figure 2). The data was distributed in the Geographic Lookup Table file (GLT) compressed format, and the uncompressed flightline images ranged in size from 0.36 GB to 44 GB. The total image dataset as delivered was 530 GB.

HyMap images were converted to apparent surface reflectance using HyCorr software. The processing consists of two levels, the simpler level is essentially compatible with ATREM3 (Atmospheric Removal program) processing. This is followed by an addition level of spectral enhancement using an EFFORT polishing pass to remove systematic ATREM3 artifact errors. The derived surface reflectance image was scaled by 10,000. Preliminary geocorrection of the flightlines was performed by HyVista with proprietary software that used sensor position and orientation data collected simultaneously with the image to calculate the position of each pixel in the image.

In addition to the hyperspectral imagery, CSTARS received a set of 1-foot color orthophotographs acquired in May 2003 by the Department of Water Resources covering the majority of Priority 1.

Improving the Spatial Accuracy of the Imagery

Comparing the Hymap imagery to the 1-foot color photographs, the image pixels were found to be 3-10 m from their true spatial location. To improve the spatial accuracy of the flightlines, we collected geographic correction points (GCP's) between the Hymap imagery and the 1-foot orthophotos, and warped each flightline of Priority 1. GCP's were spaced evenly throughout the image to ensure a more accurate geocorrection and registration of the warp image. The accuracy of these tie points was assessed by the Root Mean Square Error (RMSE), which is a statistical assessment of the accuracy of each GCP in relation to all others. Error values less than 1.5 (i.e., within the 1.5 m of the orthophoto positions) were considered acceptable (Appendix, Table 2). Visual assessment of the Hymap images after registration was performed to validate the efficacy of our techniques (Figure 6).

Efficient Extraction of Pixel Information from Field Data

Given the large volume of data acquired in this study it was necessary for us to develop specific tools for handling and analyzing the imagery in addition to those available in standard image processing software packages. As part of this study we developed a computer algorithm for fast, selective pixel extraction from remotely sensed data called STARSPAN. It works on two basic inputs; 1) raster data – one or multiple raster files from which the pixel data will be extracted, and 2) vector data – a file defining a set of features to be extracted from the raster data, such as an ArcMap shapefile. Extracted pixel data are stored in CSV files, a common text-based format that can be imported into standard applications for further analysis. Other functionalities of STARSPAN include: the generation of band specific statistics for specified features, the generation of paired pixel band values for subsequent calculation of calibration coefficients, update options, and generation of output data in ENVI formats. For more information, including installation instructions and a user manual, please visit:

Geocorrection of Flightlines

Hymap Imagery True Color Mosaics Post-Geocorrection



Hymap Imagery True Color Mosaics Pre-Geocorrection



0 0.5 1 2
Kilometers
UTM Zone 10N, NAD 83

Analysis conducted by CSTARS
University of California, Davis



Figure 6

<http://starspan.casil.ucdavis.edu>. This tool allowed us to rapidly test a variety of classification techniques without having to process each of the full flightlines.

Generating a Water Mask

To limit the image analysis and classification to specific areas within the Priority 1, we created a mask, or region to be analyzed, for just the water courses. Masking to include only the lakes, channels, and rivers improves the species determination by limiting the range of spectra analyzed and eliminating similar spectral classes that would falsely identify vegetation outside the waterways. The location and size was used to select the channels and ditches. As a general rule, channels and ditches with widths greater than 30 m (98 feet) were included, however, large water bodies that were not connected to the main channels, such as community lakes, were not included. At the extreme east and west edges of Priority 1, frequently flooded pastures and other wetlands are extensive, but cannot be navigated, even with a canoe, and were therefore, not included.

The mask was digitized using the US Bureau of Reclamation GIS layer of waterways “Levee_24_CA”, which provided a good quality line map containing the levees digitized from 1:24,000 maps. In ArcGIS, the lines were edited over a color infrared composite of the flightline images, since the levee lines fit the water boundary shape, but extended too far from the water’s edge. The lines were shifted, dissected, and additional lines digitized to fit the water boundary more tightly. For most of the levee channels, the lines were placed at the color change observed in the imagery, possibly due to the rock rip-rap and soil, slightly away from the water on the levee. In other areas, the distance from the water body was approximated based on tonal changes that expressed moist soil or vigorous vegetation from available open water in summer. After visually inspecting the digitized boundary over all the images, the polygon map was converted to a tiff image to generate the mask used in the image processing software. One additional water mask was created for the Jones Tract area owing to complications in classifying this large expanse of shallow water with detectable vegetation (i.e., fields) below the surface.

Classification Methodology for Detecting Target Species

Based on the successful performance of a technique called Spectral Mixture Analysis (SMA) in the pilot study (Mulitsch and Ustin, 2003), the current study also utilized this technique. SMA is particularly appropriate for hyperspectral imagery, and has been applied to the decomposition of mixed pixels in remotely sensed data for over 30 years (Hortwitz et al. 1975). SMA assumes that each pixel spectrum is a linear combination of spectrally distinct components, called endmembers. SMA employs least squares fitting to derive estimates of fractional composition, which are generally displayed in images as linear stretch gray levels to estimate abundances of endmembers. Each fractional image provides a subpixel estimate of both the relative endmember abundances as well as the spatial distribution of the endmember (Smith et al. 1990, Adams et al. 1995). The root mean square error (RMSE) is calculated to estimate the goodness of fit of an endmember model for mixed pixels. The portion of RMSE image having high values indicates a need to add one or more endmembers or change endmembers to improve the model. SMA has proven to be a robust technique and used to identify dynamic ecosystem changes (Adams et al. 1993, Roberts et al. 1997, Ustin et al. 1996, 1998), to map the distribution

of wetland species (Zhang et al. 1997, Sanderson et al. 1998), and for environmental monitoring (Riano et al. 2002, Small 2001, 2002, Elmore et al. 2000).

Based on an inspection of the endmembers of the target species from locations identified in fieldwork, we combined pennywort and water primrose, and tule and cattail, owing to the similarity of their spectral reflectances. Ultimately, we used six image endmembers to perform the SMA; Brazilian waterweed, water hyacinth, pennywort/water primrose, tule/cattail, water, and turbid water to achieve an optimal unmixed image (Figure 7).

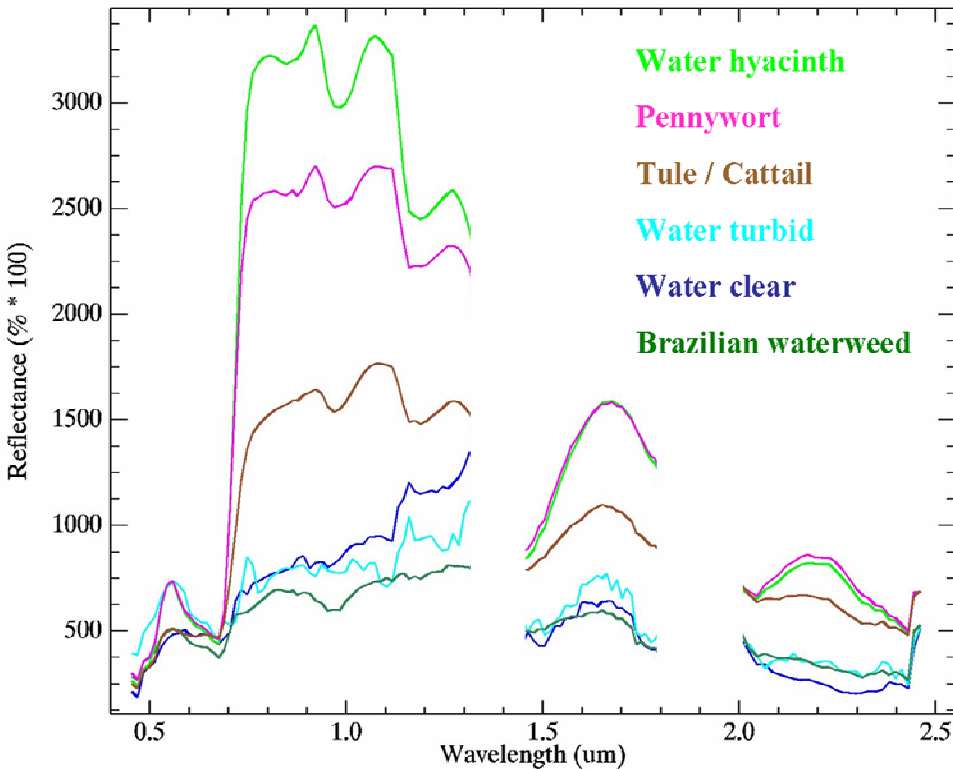


Figure 7. Representative spectra for the identified target species.

Classification of Target Species at the Flightline Scale

The classification of each flightline was performed using a manual decision tree. A series of thresholds were developed to isolate Brazilian waterweed and water hyacinth pixels. For Brazilian waterweed (Figure 8) the first step applied the water mask and ask whether the pixel was within the river channel, if yes, the second step assessed whether the pixel had a value greater than the identified threshold for Brazilian waterweed in the spectrally unmixed image. These thresholds were determined on a

Decision Tree Used to Classify Brazilian waterweed (*Egeria densa*)
from Hymap Hyperspectral Imagery

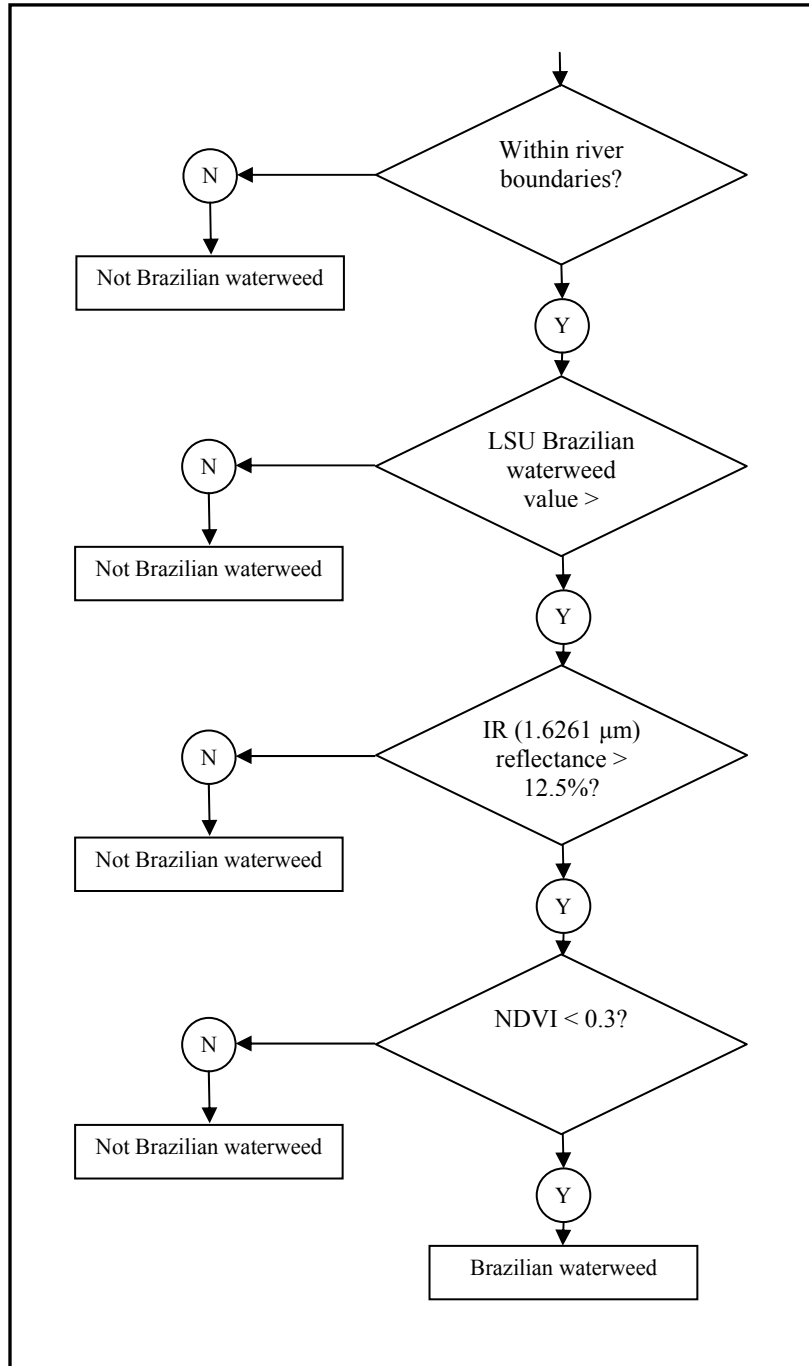


Figure 8

Decision Tree Used to Classify Water Hyacinth (*Eichhornia crassipes*) from Hymap Hyperspectral Imagery

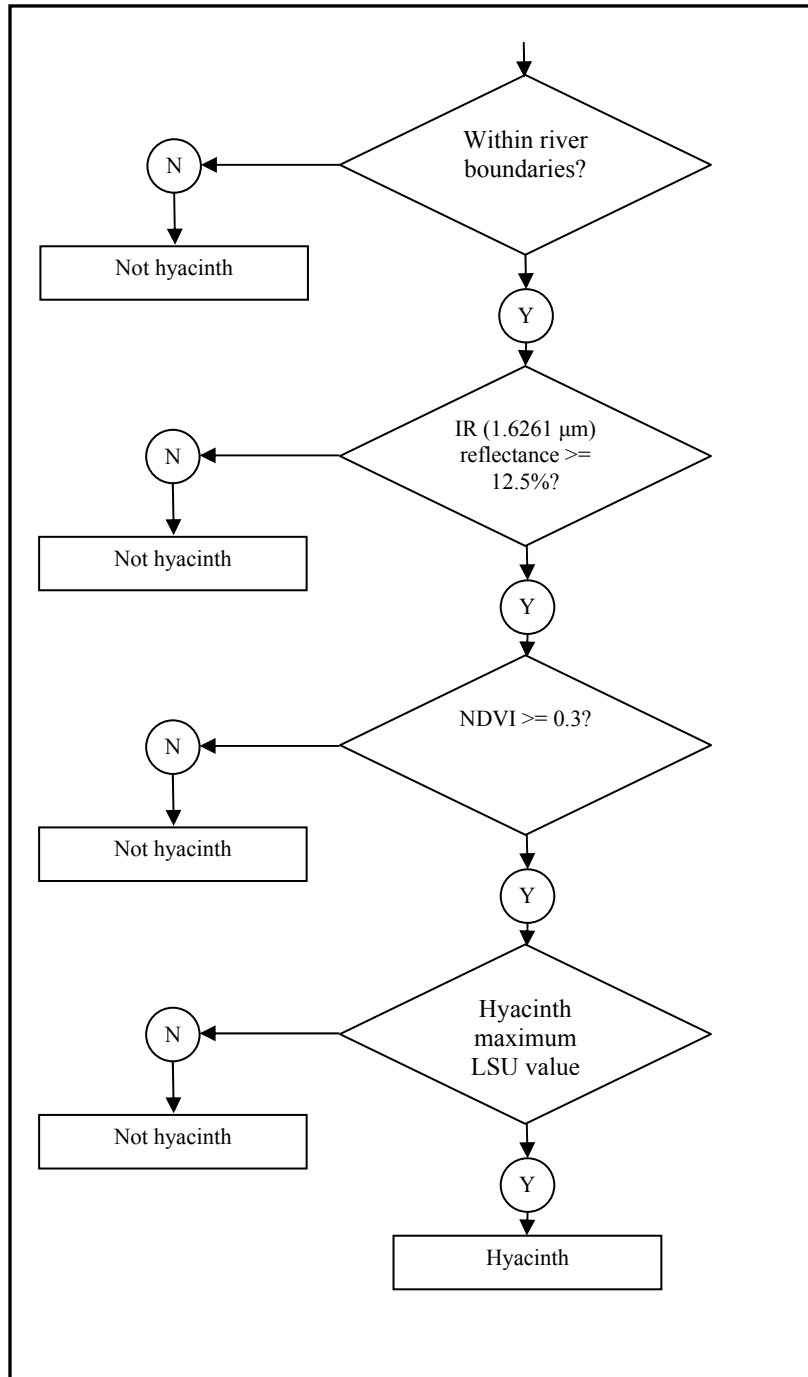


Figure 9

flightline by flightline basis using a combination of visual interpretation and statistical exploration of the field data. If the pixel value was greater than the threshold, the third step determined if the threshold value for the 1.6261 μm infrared band for that pixel was greater than 12.5%. The threshold was again identified through visual inspection of the imagery combined with information about the conditions from the field GPS points. The final step sought to distinguish Brazilian waterweed from other emerged vegetation based on selecting pixels with a Normalized Difference Vegetation Index (NDVI) <0.3 . We refer throughout this report to classifying Brazilian waterweed, since 80% or more of the SAV in the Sacramento-San Joaquin Delta is Brazilian waterweed (Julie Owens, CDBW, personnel communication), however, the mapping effort likely captures other submerged aquatic vegetation.

A different decision tree approach was utilized for classifying water hyacinth (Figure 9). In contrast to Brazilian waterweed, water hyacinth was characterized as having a high infrared reflectance ($\geq 12.5\%$ in band 1.6261 μm) and high NDVI (≥ 0.3). The high values occur because of strong absorption in the red spectral region and reflection in the infrared by the plant cells of water hyacinth. Water hyacinth had the highest endmember value among the two other major emergent vegetation classes (pennywort/water primrose and tule/cattail).

The thresholds for each target species were selected for each flightline based on our interpretation of image characteristics and the field data. Consequently, for this preliminary report, we restricted our classification to the 32 flightlines that contained the GPS points, with flightline number 23 as our northern boundary and flightline number 55 as the southern boundary.

Classification of Target Species at the Priority Site Scale

Based on recommendations from the Department of Boating and Waterways, six priority sites (three containing Brazilian waterweed and water hyacinth and three containing only water hyacinth) were selected for a finer-scale classification. One additional site, Seven Mile Slough, was classified to investigate the ability to distinguish Brazilian waterweed from different submerged aquatic species.

Classifying the target species at this scale involved a slightly different method than the decision tree approach used for the full flightlines. At the priority site scale, issues such as tidal effects and differences in water appearance were minimized, so specific endmembers for Brazilian waterweed and water hyacinth could be selected and iteratively fine-tuned after each classification was performed.

SMA, as with other image processing routines, produces rule images. These images have the same x, y configuration as the original images, but contain a new “band” for each species class. The values within each pixel are the numeric quality of the fit of the pixel spectrum to the spectral representative for the species. Generating rule images allows classing at various thresholds, or classifications performed at a later time, without the computing costs of reanalysis of the mixture modeling. In this case, pixels were classed based on the dominant species, as determined by the comparative fit among the rule bands for the species classes.

For each of the six priority sites the linear unmixed image was converted from the rule image to a classed image based on the dominant species. The sum of the pixels identified for each species, or mixture, was multiplied by the 9 m^2 per pixel to determine the area of the species. The acres of the species or mixtures are derived by converting the square meter values.

Assigning tide and date information

Based on our field observations of the effect of varying tide heights on submerged aquatic vegetation, we tagged accurate tide information to each flightline, to each priority site, and to the field GPS data. Ultimately, this information may aid in explaining inconsistencies in vegetation mapping accuracy. Information for tide monitoring stations throughout the Delta were acquired through the WWW Tide and Current Predictor at <http://tbone.biol.sc.edu/tide> and <http://www.sailwx.info/tides/index.html>, both of which utilize the accurate XTide software to predict tide height information, forward and backward in time. Further information on the XTide predictor is available at <http://www.flaterco.com/xtide/> as well as a downloadable version of the software, with a public license. The tide predictions are based on accurately assessing harmonic constants from a large number of tide stations in the Delta. Linking the ArcMap display of tide stations in the Delta to a database of the html interfaces for XTide proved to be a quick method for accessing this information at specific locations (Figure 10).

Combining Hymap Flightlines with Tide Data

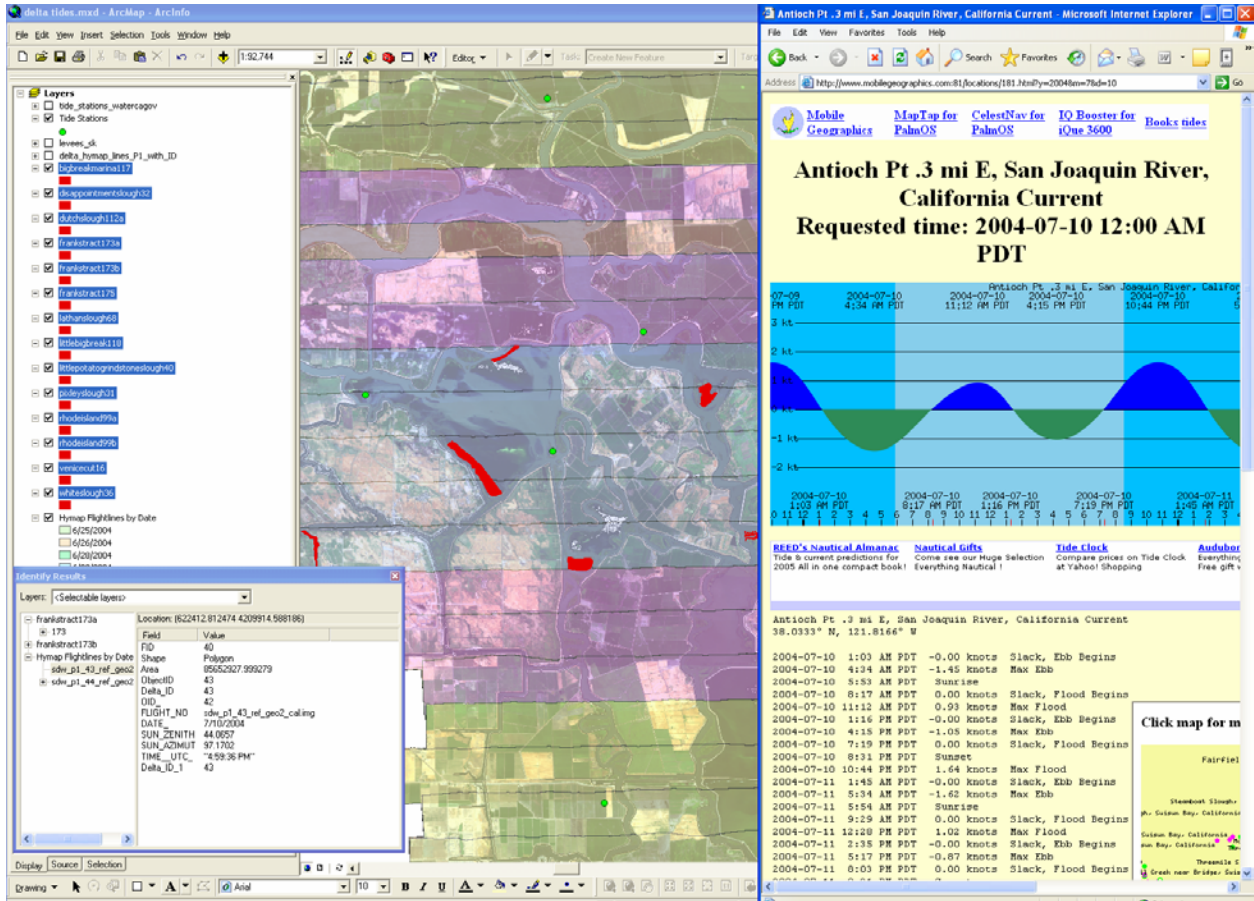


Figure 10

RESULTS

Fieldwork Data

All GPS data files were differentially corrected using Pathfinder Office (version 3, Trimble Navigation Limited, Sunnyvale, California) and exported as ArcMap shapefiles (in the Universal Transverse Mercator, zone 10, North American Datum 1983 projection). The majority of the field work was collected in the Priority 1 region (2,103 GPS points), although one team consisting of an airboat and two researchers spent three days collecting data on the Merced and Tuolumne Rivers in Priority 2 (261 GPS points).

The GPS points were screened in three ways. First, the attributes of the GPS points were checked and edited (e.g., points without an associated percent cover were deleted). Second, since many of the positions were recorded with a distance and direction offset from the boat, the GPS points were overlain on the imagery to ensure that all submerged and emerged vegetation fell within the water channels and all levee species aligned on the banks of the channels. Third, to further refine the field data we viewed it in hyperspectral space by extracting the 126 bands for each GPS point using STARSPAN. We performed a multivariate outlier analysis for all the GPS points by species, using the 117 useful bands of hyperspectral imagery associated with each point. This procedure estimates the mean, standard deviation and correlation matrix of the data (jackknife technique) and extreme outliers were then identified and removed. Polygons acquired in the field were not screened, but serve to provide additional reference data for assessing the validation of the classification.

Over 200 GPS points from Priority 1 were discarded through this screening process (Appendix, Table 1). Of the submerged aquatic species, Brazilian waterweed had the greatest number of GPS points acquired (432 GPS points). In addition, a substantial number of other submerged aquatic species, such as watermilfoil (*Myriophyllum spicatum*), coontail (*Ceratophyllum demersum*), cabomba (*Cabomba caroliniana*), curly leaf pondweed (*Potamogeton crispus*), and American pondweed (*Potamogeton nodosus*) were acquired, which will be utilized to test the ability of distinguishing Brazilian waterweed from the other submerged species. Of the emergent species, water hyacinth had 262 GPS points acquired, with slightly fewer points for pennywort (201) and less for water primrose (71). The main levee species recorded were tule (259 GPS points) and varying numbers for other levee species. Twelve pepperweed points and no purple loosestrife points were acquired in Priority 1.

Classification of Target Species at the Flightline Scale

A decision tree approach was used for classifying Brazilian waterweed and water hyacinth which involved assessing thresholds for each species from; spectrally unmixed images, one infrared waveband, and an NDVI value. An accuracy assessment of the two target species between flightlines 23 and 55 showed that 29% of the field collected Brazilian waterweed locations were correctly identified by the classification and 65% of the water hyacinth points were correctly classified (Figure 11). The most notable concentration of Brazilian waterweed is in the Franks Tract area along the south-western

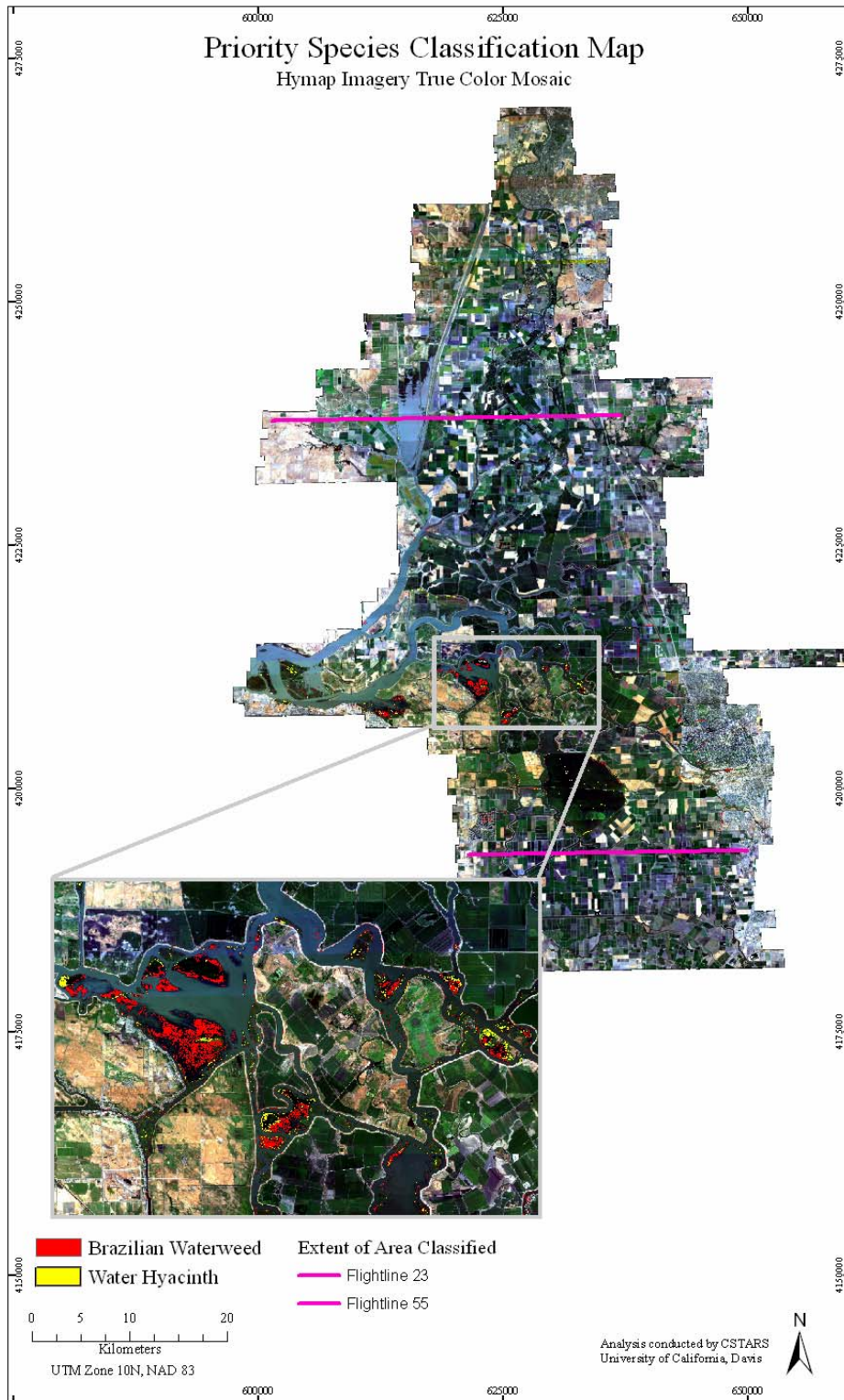


Figure 11

side and at Big Break to the west of this area. The inset map of Figure 11 also clearly shows the concentration of this species around Rhode Island and Venice Cut, and further to the east along Disappointment Slough. We feel most confident about the results acquired in the western and central portions of the study area, while the slower moving waterways and sloughs in the eastern side of the study site presented significant challenges for mapping Brazilian waterweed (see Discussion). The difficulties encountered in mapping at this coarse scale and the resultant accuracy are in contrast to results from mapping Brazilian waterweed at the finer scale of the priority sites which, according to field data and observations from fieldwork, appear to be much more accurate. Although the overall mapping accuracy for Brazilian waterweed is relatively low, we note that the results show distinct patterns of infestation which provide a useful guide for management decisions relating to control efforts and serves as a baseline dataset to monitor future expansion.

The accuracy of classifying water hyacinth over these 32 flightlines is much higher. The areas of infestation are much smaller in size than those of Brazilian waterweed often concentrated in stable waters around islands, and thus they are difficult to decipher at the resolution of Figure 11.

Classification of Target Species at the Priority Site Scale

1. Rhode Island

Flown on 7-10-2004

Tidal height at time of over flight: ~1.4 ft

Flightline 45b

Rhode Island is a very sheltered site. Only one small channel opening provides access to the site from a channel on the western side. The outer areas of the site are dominated by emergent vegetation. Figures 12a-d shows the gray scale fraction images for all unmixed species. Areas that are bright white correspond to dense areas of the target species. In Figure 12a and 12b, tule/cattail is depicted as green dots, and pennywort/primrose as cyan dots at the outer edges of the site. A couple of areas of water hyacinth are located on the inside of Rhode Island, as seen in Figures 12c and 13. The yellow dots depict field locations of known water hyacinth, corresponding to the bright areas in the fractional image. To assess the accuracy of our classification we overlaid the field collected GPS points on the fractional image and calculated the number of GPS points that fell on pixels with a value ≥ 0.5 , i.e. where the target species composed the majority of the pixel. For water hyacinth three out four GPS points collected in the field were correctly mapped using SMA.

Brazilian waterweed is highly prevalent in the sheltered areas of this site. As seen in Figures 12d and 13, the entire middle section of the site is mapped as Brazilian waterweed. The GPS points of this submerged weed, denoted by red dots, are located in bright white pixels, indicating that SMA successfully mapped Brazilian waterweed at this site. Seven out of nine Brazilian waterweed points were mapped correctly. The total acreage for the weeds at this site was 66.4 acres for Brazilian waterweed and 7.8 acres for water hyacinth (Table 1).

2. Venice Cut

Flown on 7-10-2004

Tidal height at time of over flight: ~0.6 ft

Flightline 42

Venice Cut is sheltered on three sides by emergent plants along the outer shoreline edges and the interior is dominated by SAV. Figures 14a and 14b are the gray scale fractional images with the target species water hyacinth and Brazilian waterweed in bright white pixels. Figure 15 shows this distribution overlaid on the true color mosaic. GPS field collected points, shown as red and yellow dots, are located in bright white pixels, indicating that SMA successfully mapped both the water hyacinth along the outer edges and SAV in the middle areas. All eighteen Brazilian waterweed points and five of seven water hyacinth points had values ≥ 0.05 in their respective fractional images and were thus correctly identified. The total acreage for the weeds at this site was 72.5 acres for Brazilian waterweed and 5.9 acres for water hyacinth (Table 1).

3. Pixley Slough

Flown on 7-10-2004

Tidal height at time of over flight: ~0.5 ft

Flightline 42

Pixley Slough is composed of two narrow channels running west to east that are joined together by a strait. This site is located on the far eastern side of the Delta study area and characterized by riparian vegetation. As seen in Figures 16a and 17, Brazilian waterweed is very distinct along the upper channel and along the narrow straits to the east. This corresponds well with locations of the aquatic weeds that were collected in the field, seen as red dots on the gray scale image. Fourteen of sixteen Brazilian waterweed points fell onto pixels which had values of ≥ 0.05 and were therefore mapped correctly. Brighter pixels of water hyacinth are located in the northern channel and along the levee banks of the southern channel (Figures 16b and 17). Mapped areas corresponded well with the GPS locations for the weed. Six of nine water hyacinth points overlaid with the pixels dominated by water hyacinth in the SMA fractional image. The total acreage for the weeds at this site was 16.5 acres for Brazilian waterweed and 3.7 acres for water hyacinth (Table 1).

4. Dutch Slough

Flown on 7-10-2004

Tidal height at time of over flight: ~2.3 ft

Flightline 45b

Dutch Slough channel narrows into a very small area surrounded by riprap levees. As seen in Figures 18a and 19 the majority of the area is dominated by Brazilian waterweed, the southern section most heavily dominated as depicted in white pixels. All eight Brazilian waterweed field points overlaid correctly on the bright white pixels and thus with pixels dominated with that species in the SMA fractional image. Water hyacinth mats are seen in Figures 18b and 19 along the outer edges of Dutch Slough. Three of four water hyacinth points were mapped correctly. The total acreage for the weeds at this site was 22.7 acres for Brazilian waterweed and 1.0 acres for water hyacinth (Table 1).

5. Disappointment Slough

Flown on 7-10-2004

Tidal height at time of over flight: ~0.5 ft

Flightline 42

Disappointment Slough is a control site located in the eastern part of the Delta. In Figures 20a and 21, Brazilian waterweed is seen as bright white pixels in the curved channels of the slough. A couple of large water hyacinth mats can be seen in the eastern section of the slough (Figures 20b and 21). GPS points of the Brazilian waterweed and water hyacinth correspond with mapped areas of the weeds. All seven Brazilian waterweed GPS points were mapped correctly as were the two water hyacinth points. The total acreage for the weeds at this site was 70.8 acres for Brazilian waterweed and 4.5 acres for Water hyacinth (Table 1).

6. Latham Slough

Flown on 7-10-2004

Tidal height at time of over flight: ~1.0 ft

Flightline 45b

Latham Slough is another control site. As seen in Figure 22a and 23, the waterways surrounding the five fingers of emergent vegetation are highly infested with the Brazilian waterweed. Mapped locations correspond with field collected GPS points. Nine of twelve Brazilian waterweed points were correct in that those pixels were classified as being dominated by Brazilian waterweed. The middle area is either less dominated by weeds or perhaps flooded to a greater depth. A large patch of water hyacinth is visible in Figures 22b and 23. A few smaller patches of water hyacinth are seen along the outer edges of the emergent vegetation. All five water hyacinth points overlaid correctly on pixels dominated by this species in the SMA fractional image. The patches of water hyacinth in the five fingers region were mixed with larger stands of tule and cattail. These water hyacinth mats are clearly identified among the stands of emergent vegetation in the fractional images. The total acreage for the weeds at this site was 20.1 acres for Brazilian waterweed and 53.6 acres for water hyacinth (Table 1).

7. Seven Mile Slough

Flown on 7-9-2004

Tidal height at time of overflight ~1.9 ft

Flightline 37

Seven Mile Slough, near Owl Harbor, was identified as a potential priority site in 2004 and was a priority site in 2003. This site was chosen to investigate whether other submerged aquatic vegetation can be differentiated from Brazilian waterweed. In addition to Brazilian waterweed, coontail, curly leaf pondweed, American pondweed, and watermilfoil were present at the site. American pondweed was spectrally distinct from Brazilian waterweed and the other submerged aquatic vegetation. Mats of Brazilian waterweed are seen in Figure 24a surrounding islands of tule and cattail. Three of four Brazilian waterweed points were correctly mapped in the SMA fractional images. In comparison to 2003, Brazilian waterweed was not the dominant submerged aquatic vegetation at the site, which is probably due to Diquat treatments conducted last year by CDBW.

As seen in Figure 24b, one large patch of American pondweed can be seen in the middle of the image. Our GPS points collected in the field overlaid this bright white patch and verified that this 30 m x 30 m surface patch was American pondweed. We also distinguished coontail

from Brazilian waterweed. As seen in Figure 24c, the northwest area of the slough is very dense with coontail mats. This corresponds with field collected location data that overlaid the bright white pixels. Additional coontail mats are present along the northern levees in the slough. All eight coontail points were correctly mapped, i.e., coontail was the dominant species within the pixels. These preliminary results are encouraging and we will continue to investigate whether hyperspectral imagery can be used to detect other submerged aquatic vegetation from Brazilian waterweed at this and other sites in the Delta.

Table 1. Area of target species in six priority areas identified by California Department of Boating and Waterways.

Priority Areas	Total Area		Water hyacinth			Brazilian waterweed			Tule / Cattail		
	Acres	Sq. Meters	Acres	Sq. Meters	Extent (%)	Acres	Sq. Meters	Extent (%)	Acres	Sq. Meters	Extent (%)
Disappointment Slough	115.6	467,901	4.5	18,198	3.9	70.8	286,542	61.2	34.4	139,176	29.7
Dutch Slough	25.3	102,402	1.0	3,987	3.9	22.7	91,719	89.6	0.0	36	<0.1
Latham Slough	155.7	629,937	53.6	216,729	34.4	20.1	81,333	12.9	11.6	46,836	7.4
Pixley Slough	114.3	462,708	3.7	15,066	3.3	16.5	66,915	14.5	25.9	104,922	22.7
Rhode Island	123.3	499,041	7.8	31,662	6.3	66.4	268,875	53.9	14.1	56,907	11.4
Venice Cut	249.9	1,011,456	5.9	23,976	2.4	72.5	293,553	29.0	64.1	259,398	25.6

Rhode Island

Tule/Cattail

Pennywort/Primrose

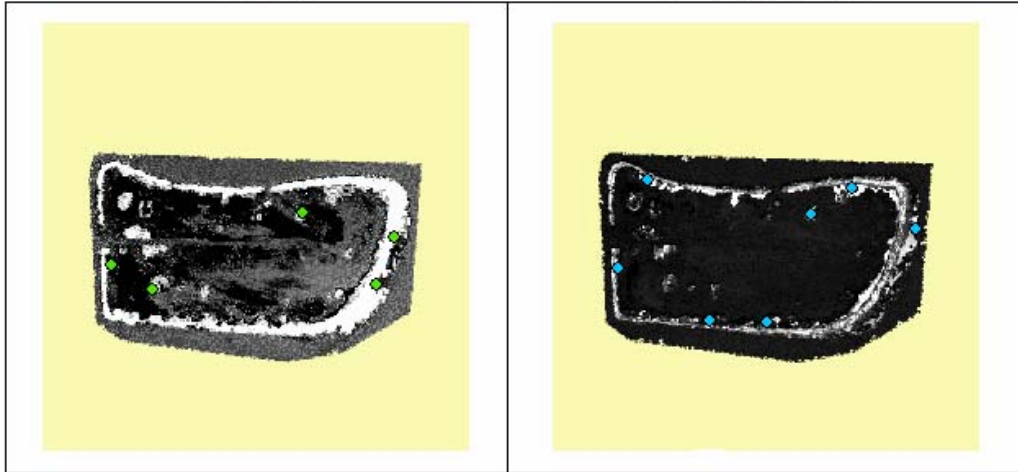


Figure 12a

Figure 12b

- Tule/Cattail
- Pennywort/Primrose
- Water hyacinth
- Brazilian waterweed

Water hyacinth

Brazilian waterweed

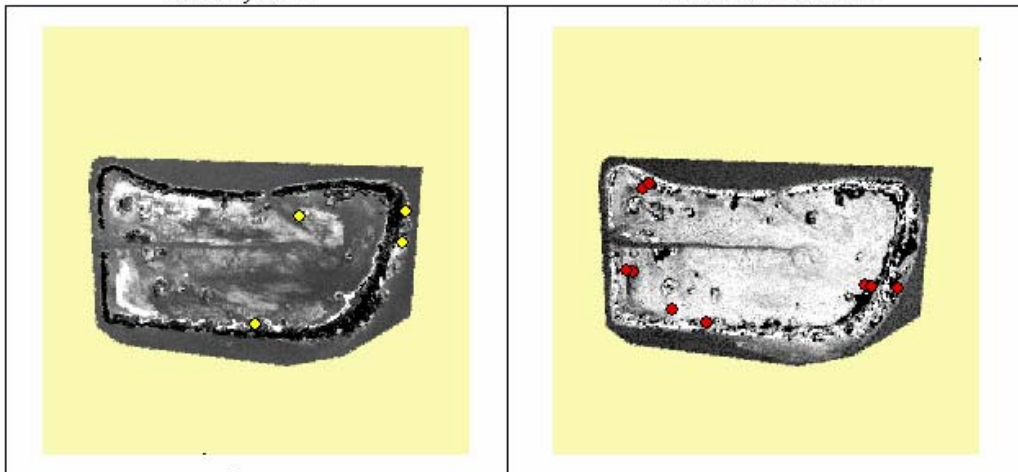


Figure 12c

Figure 12d

Analysis conducted by CSTARS
University of California, Davis

Figure 12

Classification of Target Species at Rhode Island

Hymap Imagery True Color Mosaic

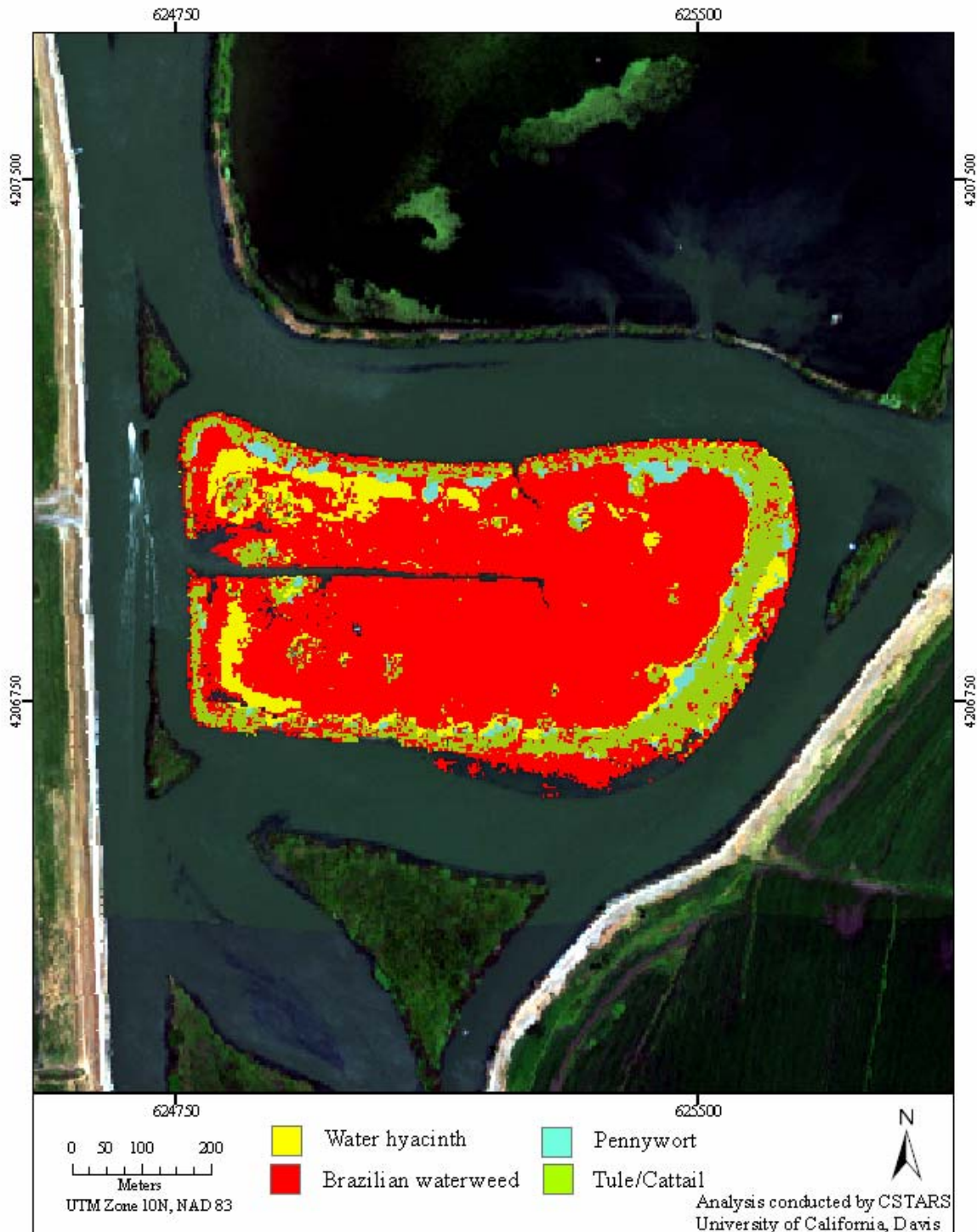


Figure 13

Venice Cut

Water Hyacinth

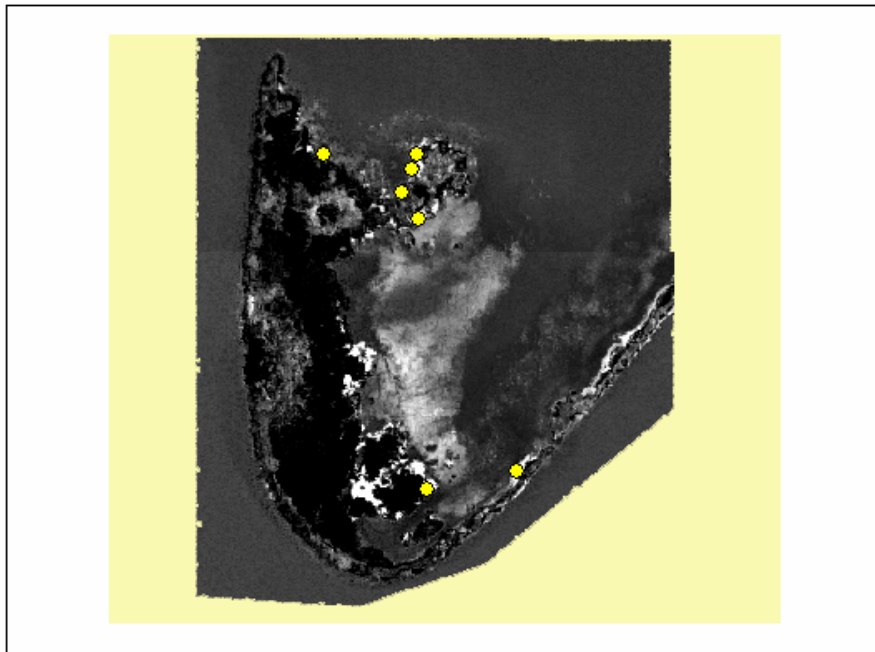


Figure 14a

● Water Hyacinth

● Brazilian waterweed

Brazilian waterweed

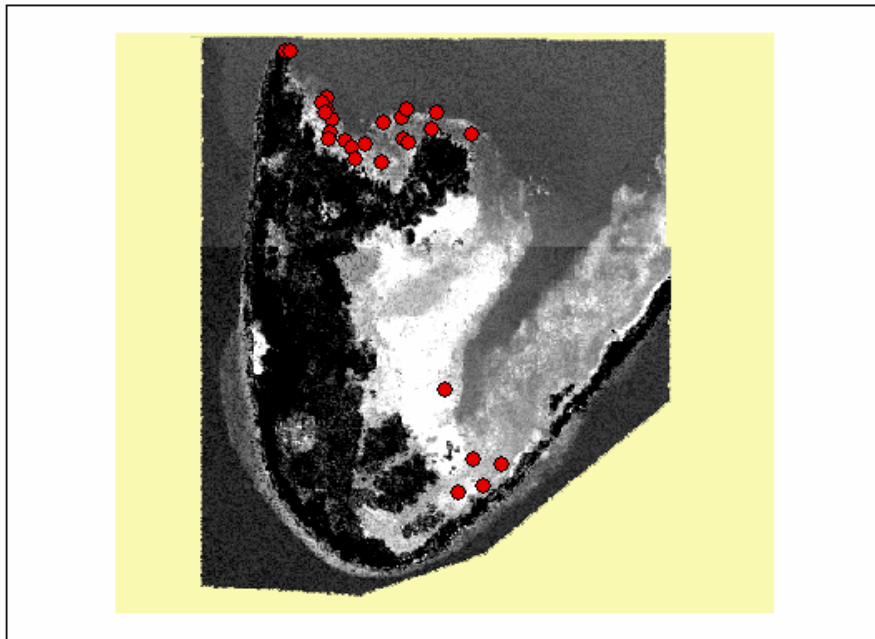


Figure 14b

Analysis conducted by CSTARs
University of California, Davis

Figure 14

Classification of Target Species at Venice Cut

Hymap Imagery True Color Mosaic

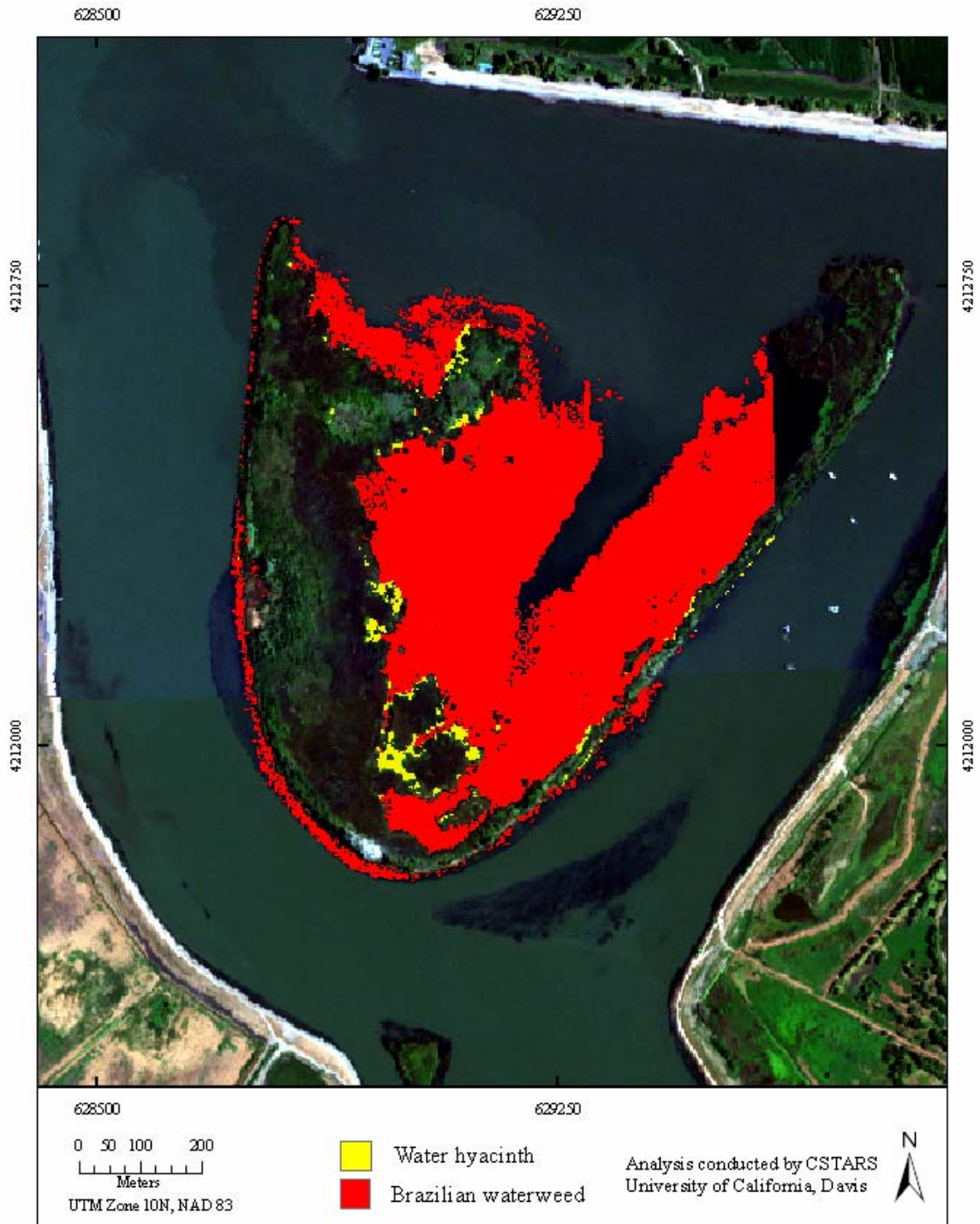


Figure 15

Pixley Slough

Water hyacinth

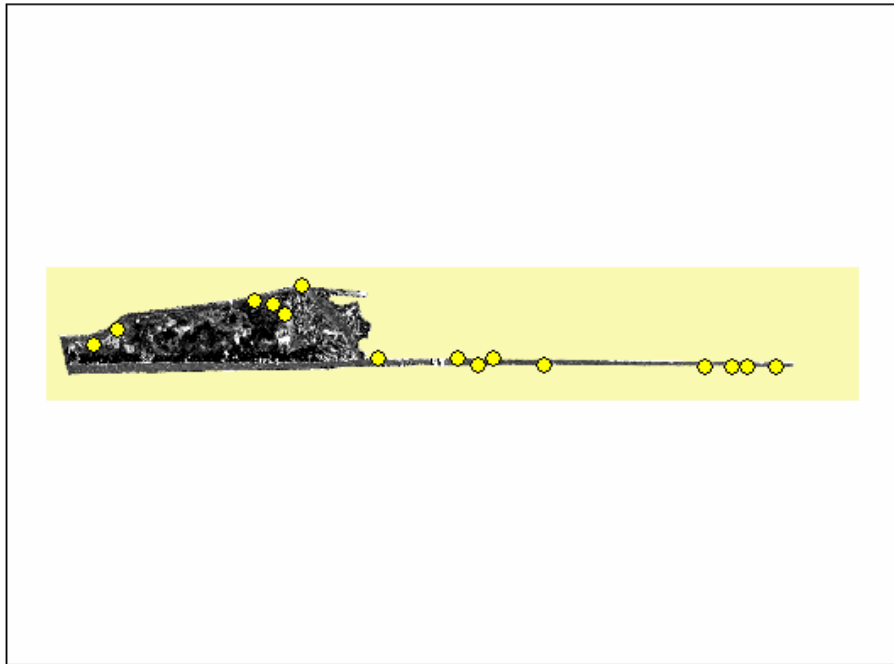


Figure 16a

- Water hyacinth
 - Brazilian waterweed
- Brazilian waterweed

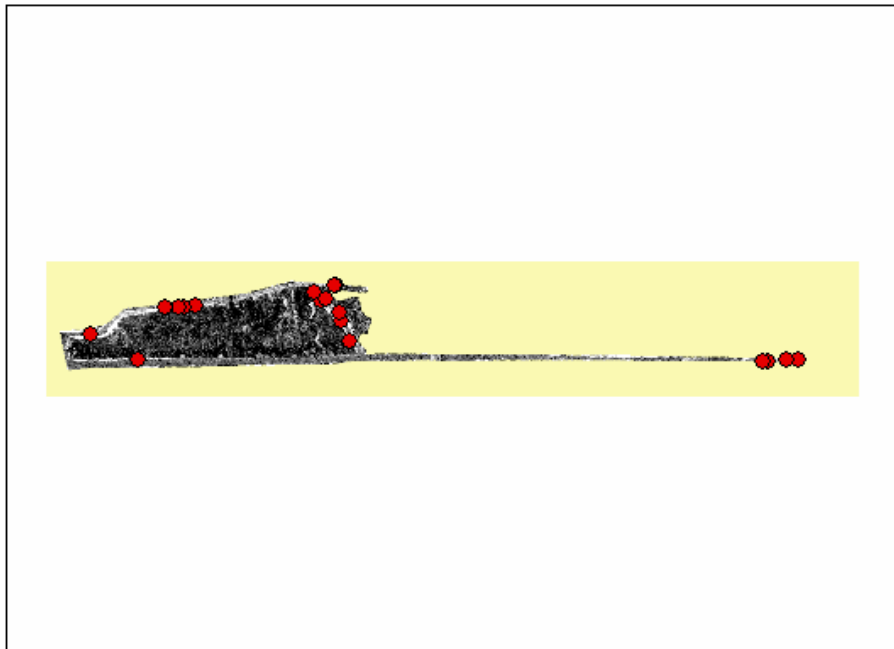


Figure 16b

Analysis conducted by CSTARS
University of California, Davis

Figure 16

Classification of Target Species at Pixley Slough

Hymap Imagery True Color Mosaic

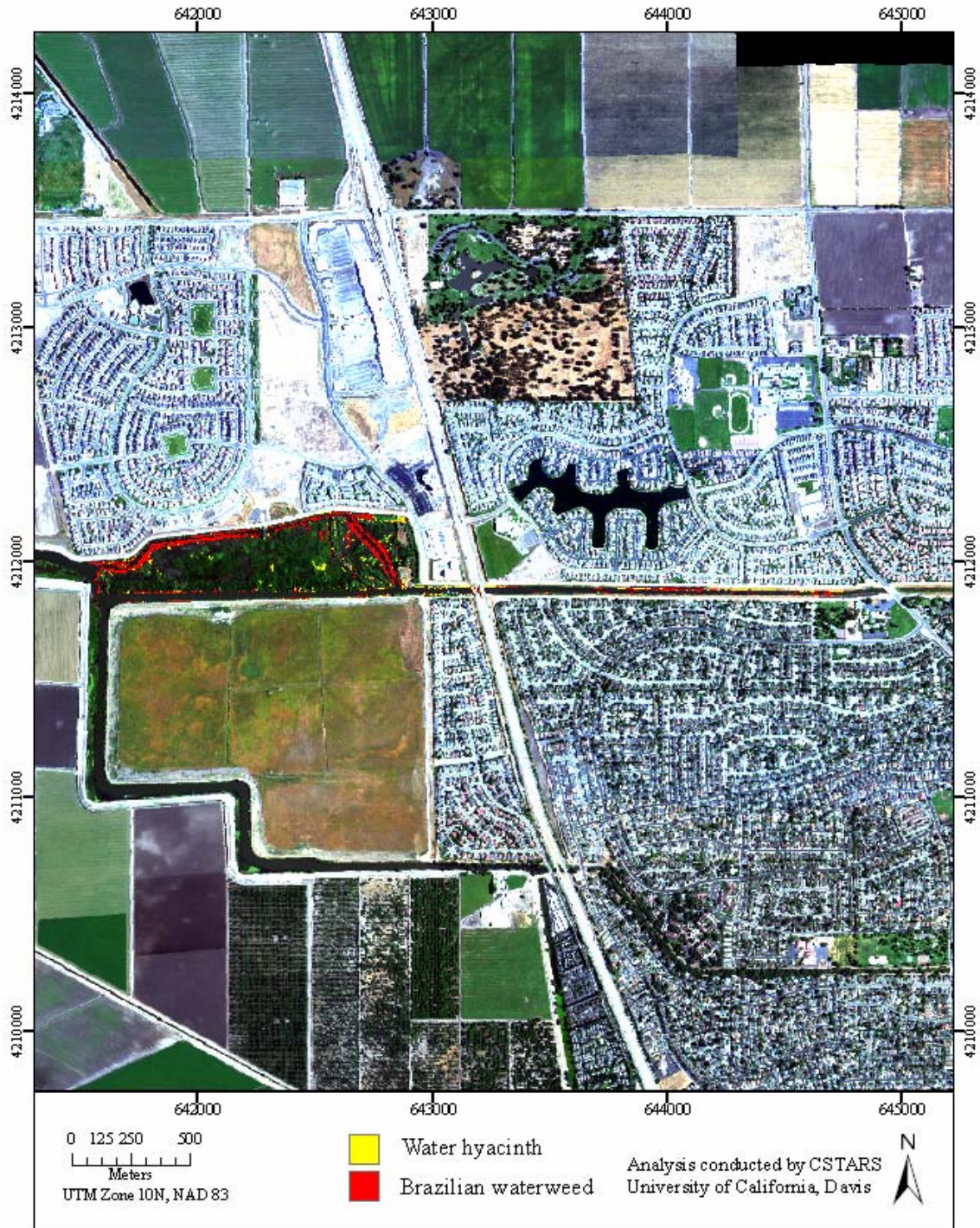


Figure 17

Dutch Slough

Water Hyacinth

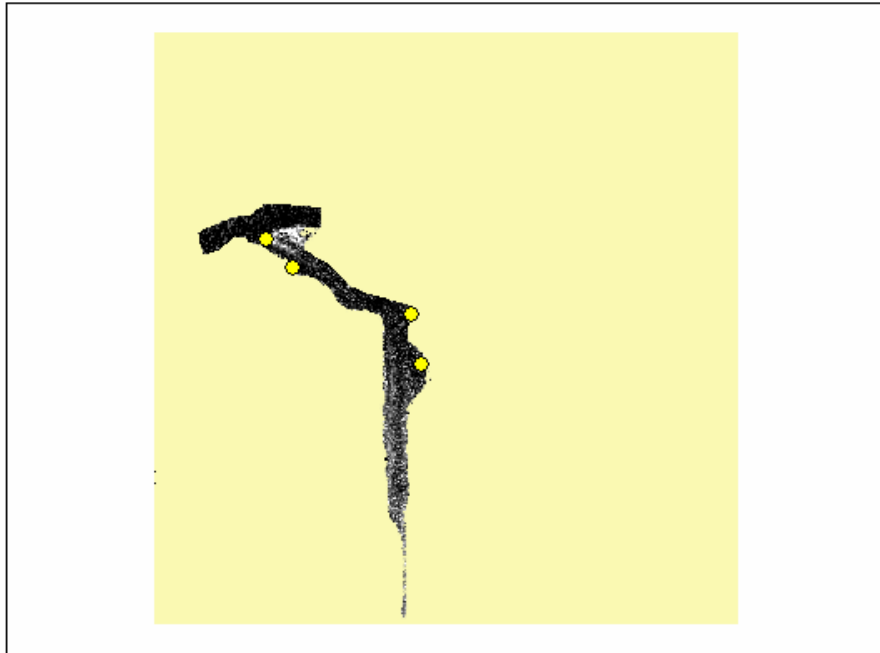


Figure 18a

- Water Hyacinth
- Brazilian waterweed

Brazilian waterweed

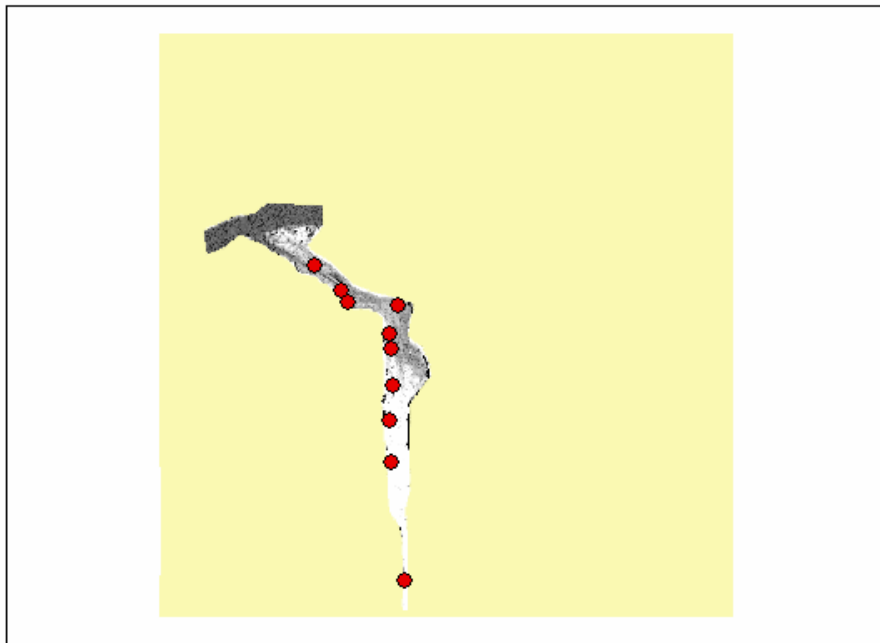


Figure 18b

Analysis conducted by CSTARS
University of California, Davis

Figure 18

Classification of Target Species at Dutch Slough

Hymap Imagery True Color Mosaic



Figure 19

Disappointment Slough

Water hyacinth

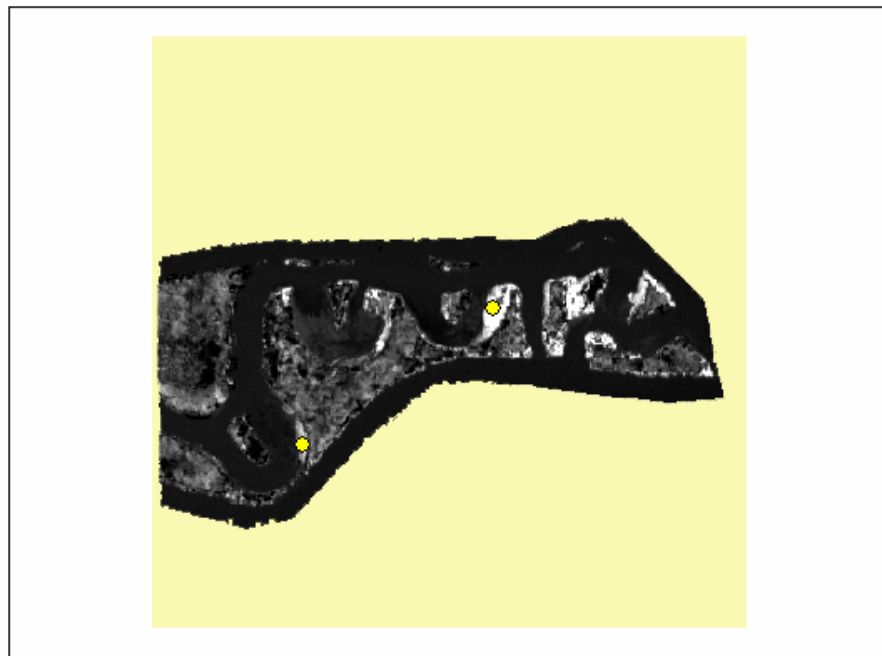


Figure 20a

- Water hyacinth
- Brazilian waterweed

Brazilian waterweed

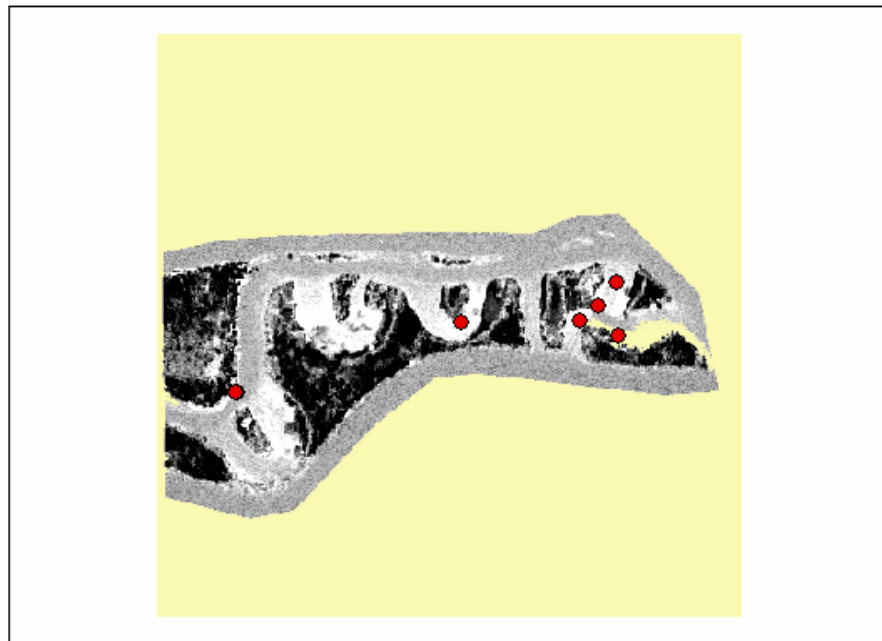


Figure 20b

Analysis conducted by CSTARS
University of California, Davis

Figure 20

Classification of Target Species at Disappointment Slough

Hymap Imagery True Color Mosaic



Figure 21

Latham Slough

Water hyacinth

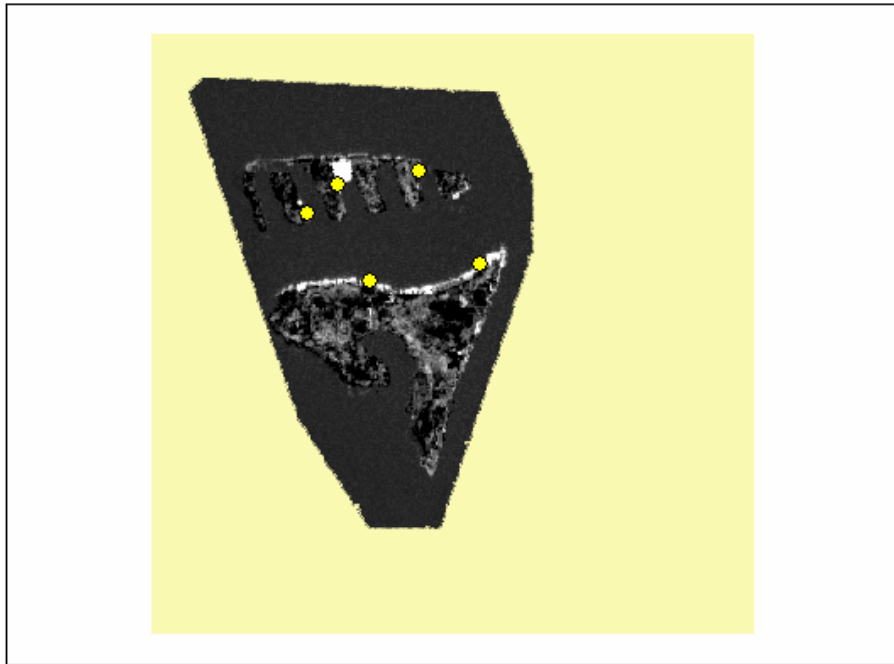


Figure 22a

- Water hyacinth
- Brazilian waterweed

Brazilian waterweed

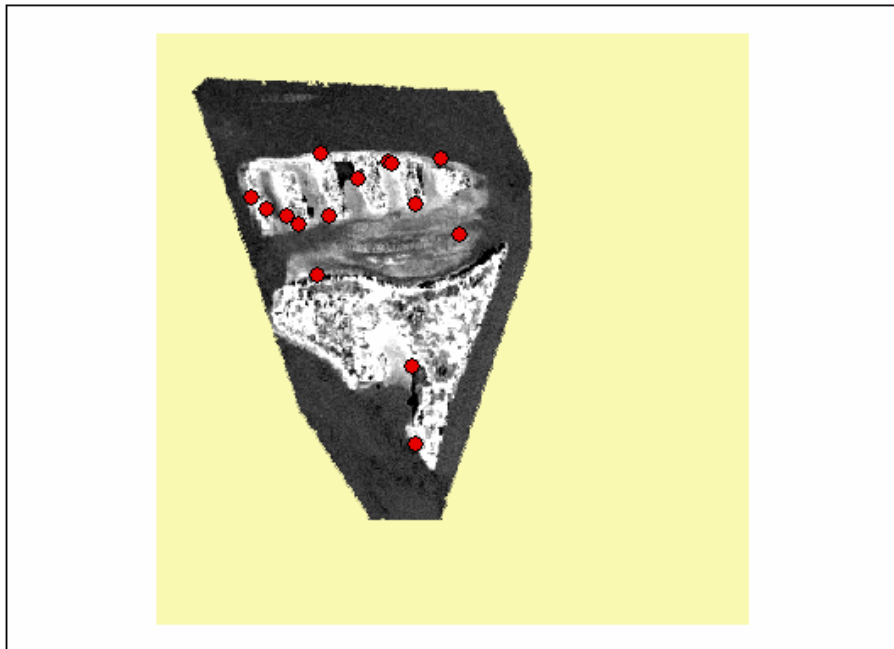


Figure 22b

Analysis conducted by CSTARS
University of California, Davis

Figure 22

Classification of Target Species at Latham Slough

Hymap Imagery True Color Mosaic

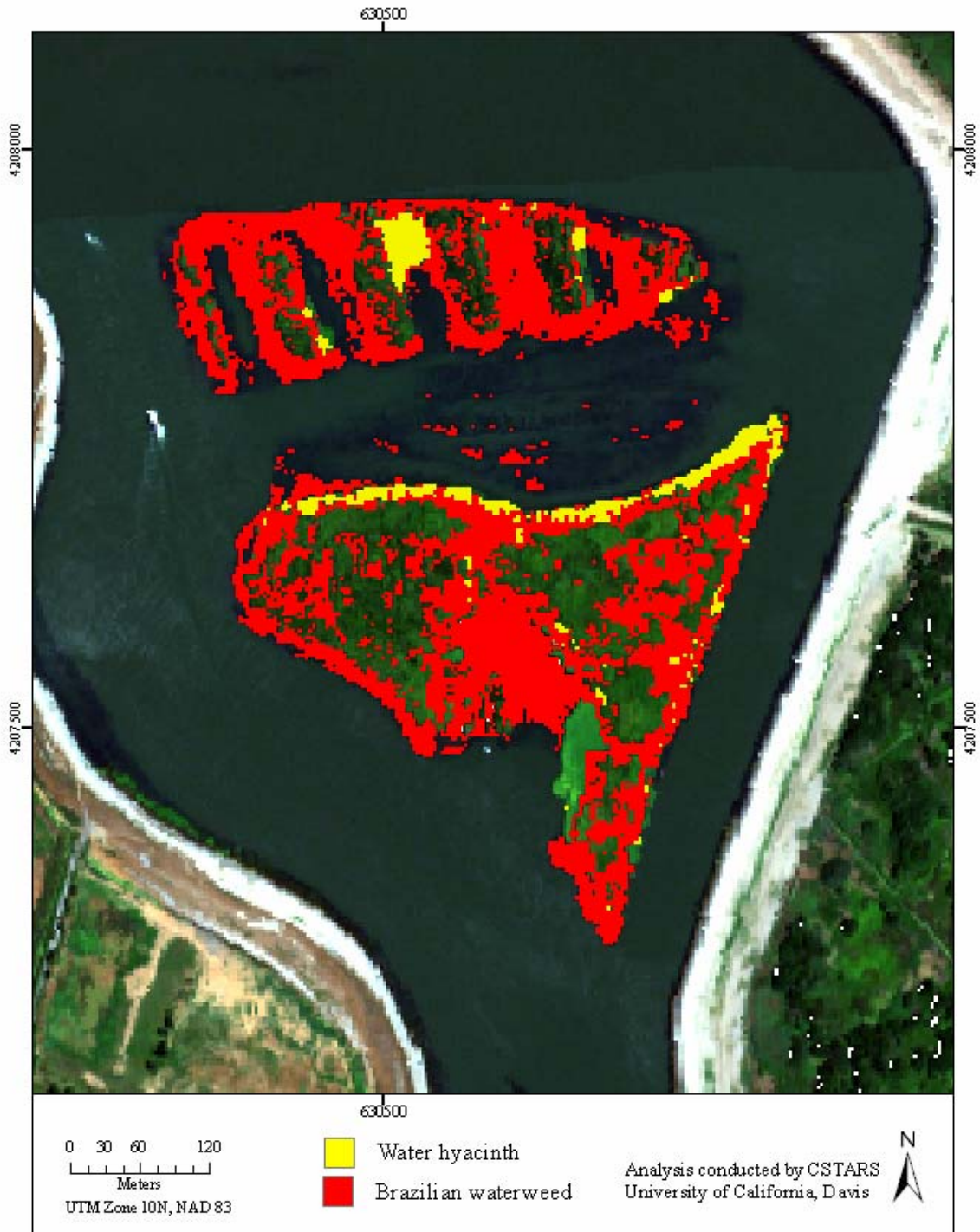
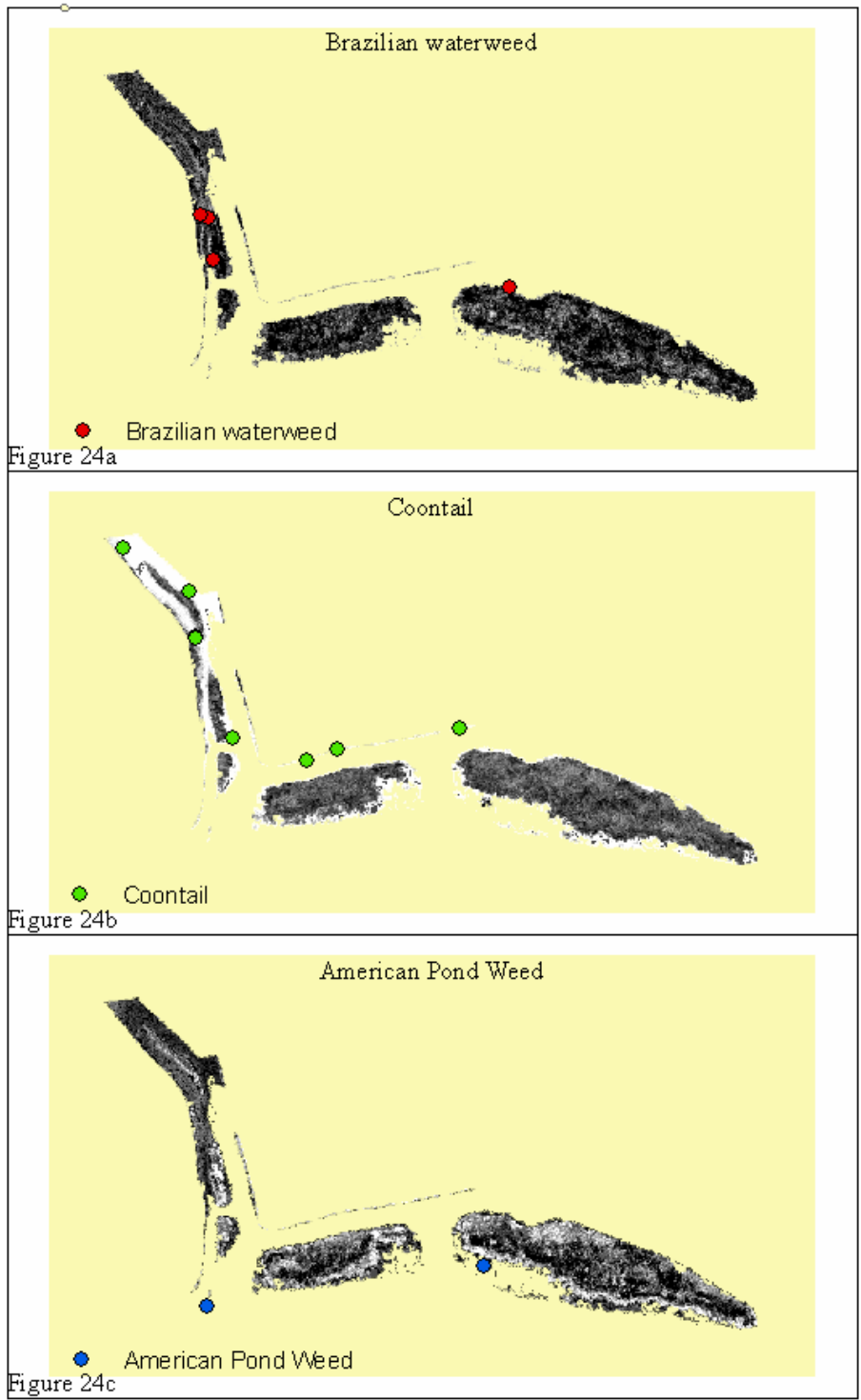


Figure 23

Seven Mile Slough



Analysis conducted by CSTARS
University of California, Davis

Figure 24

DISCUSSION

While the priority areas demonstrated that high accuracy classification is achievable, the variability in the spectra for the individual target species and water for the entire project area produced low accuracy for some species. In part, the lower accuracy was due to the broad scale of the project and the amount of imagery acquired. This project represents one of the largest airborne hyperspectral image acquisition efforts undertaken for vegetation mapping purposes. As a consequence, we encountered a number of issues that were not seen in the smaller area of 10 flightlines in the 2003 pilot project. A number of these issues are described below, followed by a discussion of potential steps to improve the accuracy of the mapping.

Field work Protocol

One of the difficulties encountered in this study stemmed from inaccuracies in the geographic positioning and species recording during the field data acquisition. The airboats were successful in driving on top of Brazilian waterweed and water hyacinth patches for accurate GPS positions, however, the motorboats were unable to enter patches of submerged vegetation without trapping weeds in the propellers. In these situations, the field crew recorded the GPS positions by estimating the distance and direction offset. Given inherent differences between observers, some GPS points may have been erroneously placed with the non-target species. The small ~3 m pixel size made precise positioning essential for successful mapping in the field and in the imagery.

In selecting representative vegetation sites, adequate cover size and density estimates are needed. The cover viewed nearly horizontally from the boats is substantially greater than when viewed from above by the airborne instrument. At the oblique angle of the observers, offsets in the sample points may have appeared greater than actual cover density.

Newer techniques such as real-time display of aerial and satellite imagery within the GPS may produce greater precision in selecting the pixels that best represent the locations being mapped in the field.

Image Georegistration

Accessible representative vegetation sites were often no larger than one pixel (3 m), therefore extremely accurate georegistration was required to properly extract the spectral data from the images. The image georegistration provided by HyVista was based on their airborne ephemeris location and plane attitude data which was not sufficiently accurate, ranging from 3 to 10 m off position. Significant effort was made to register the imagery to the 1-foot orthophotos provided by the Department of Water Resources. While this improved our positional accuracy, our classification results did improve significantly by averaging the individual pixel classifications using a 3 x 3 pixel window (median filter) over the final results. As shown in the Results section (and Appendix, Table 2), the georegistration accuracies are within a single pixel across all control points within each image, even this single pixel error can affect the accuracy of the classification. One suggestion for future data collection is to utilize GPS units that

display an existing current or recent aerial image to acquire while in the field clear and distinct feature positions for registration.

Variation in Tidal Height and Water Quality

One major problem encountered in the study this year in identifying submerged vegetation, compared to the 2003 pilot project, was the large variation in water depth due to tidal differences. The best opportunity to distinguished SAV from the surrounding water was during low tide when the tops of the vegetation are afloat on the water surface. In contrast to floating mats of water hyacinth, Brazilian waterweed and other SAV are rooted to the channel bottoms, as a consequence increased water height exposed less SAV, and the image pixel spectra was composed of a greater proportion of water spectra. Given the large number flightlines, combined with wind and cloud considerations, image acquisitions spanned almost two weeks. Our opportunity to coordinate the flight times with low tides was constrained by these and other operational conditions. In addition, the channel size and meandering patterns delay tidal movement and promote inconsistent water heights, from the western to eastern and from the northern to southern area of the study. The disparities between the tidal heights at the time the field crew collected representative sites and when the imagery was acquired added errors to acquiring representative locations.

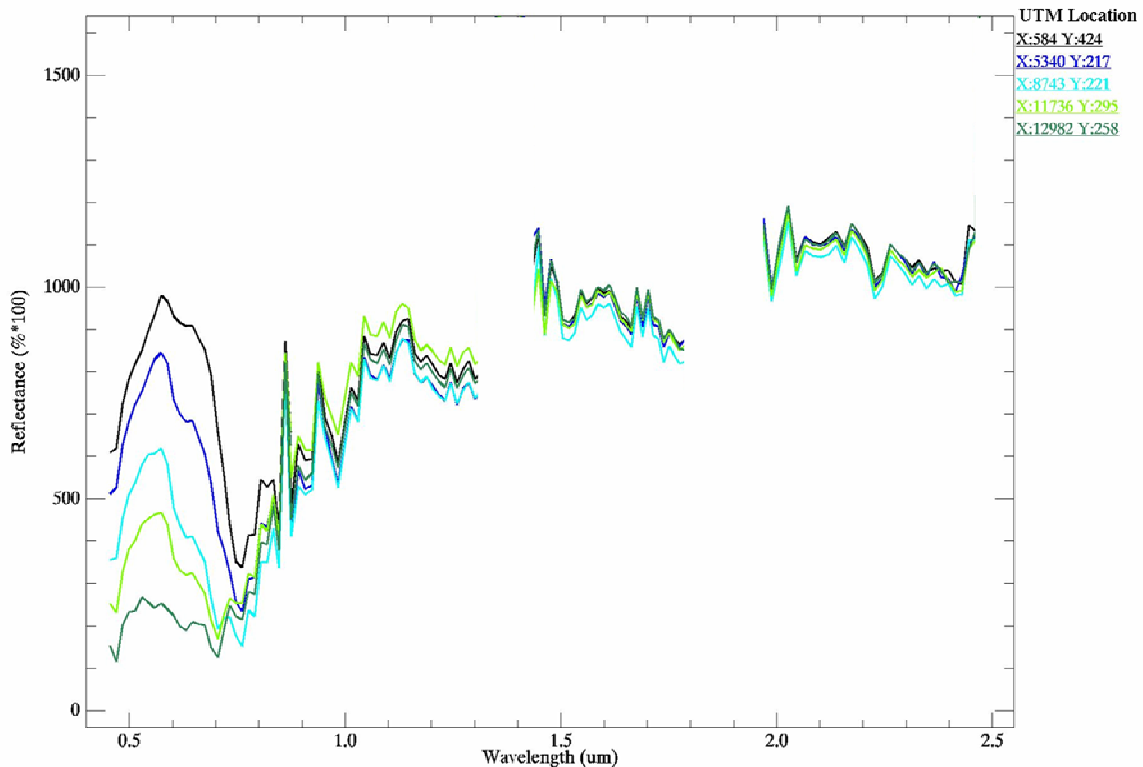


Figure 25. Flightline number 40 extracted water spectra from west to east at channel centers

The identification of Brazilian waterweed and other SAV was further complicated by the variability in water spectra due to the range of turbidity and algae within the Delta. Since turbid water spectra are similar in many wavelengths to that of submerged aquatic vegetation, it was difficult to extract adequate spectral representatives of the water. Example spectral from the middle of the Delta study area is shown in Figure 25 for the selected Universal Transverse Mercator (UTM) coordinates in the legend. These water spectra were extracted along a transect of channel centers from west to east in flightline number 40, a flightline that encompasses both the Sacramento and San Joaquin Rivers, the vegetation edges of Franks Tract and Honker Cut. In this Delta region, the western channels are cleaner, larger water bodies. The rate water is refreshed is slower in the middle and eastern region due to smaller channels. Generally the water in the northern flightlines is clear with brighter visible spectral regions due to fine clay and silt reflectance. In the southern flightlines, the water is darker in the visible region due to light absorption by a mixture of algae and other detritus. Additionally, algae in the water may be confused with the greenness of Brazilian water weed in the water, and further confounded as the amount of water in the pixel varies. The mix of green water and green vegetation is not easily discriminated.

In the future, the sub-dividing the project into smaller areas where the timing of the tide and flight can be better coordinated, and greater control on the water spectra can be achieved will improve the accuracy of the vegetation classification. Further refinement of the water spectra and vegetation spectral analysis will also improve the classification accuracy.

ReMetrix Pre-Treatment Field Support for Hyperspectral Image Classification

ReMetrix field crews collected pre-treatment surface vegetation location and measurement information by sampling eight priority sites between 25 June and 9 July 2004, prior to herbicide application. The eight sites sampled were: Dutch Slough 112, Franks Tract 173, Big Break 118, Little Potato Slough/Grindstone 40, Pixley Slough 31, Rhode Island 99, Venice Cut 16, and White Slough 36. The submerged plant samples were collected in transects across each site using a thatch rake device. They provided the vegetation information in an Excel spreadsheet. Sample positions and surface vegetation areas were mapped using GPS and supplied as ArcMap shapefiles. In addition, ReMetrix supplied tidal information for their sample collection times. By collecting data during high tide, the submerged plants are distributed as much as possible throughout the water column. During low tide, the dense plants tended to tangle into mats at the water surface, reducing the accuracy of the ReMetrix acoustic measurements.

Unfortunately, the ReMetrix requirement to collect their data at high tide is the opposite of our requirement to acquire imagery at or around low tide for species identification. Weed patches that may be present at low tide and in the imagery may not have been observed by the ReMetrix field crews.

ReMetrix supplied shapefiles for mixed vegetation patch locations of five of the eight sites. The location and identification of the patches helped confirm the presence of mixtures of submerged weeds, however, we were unable to use these mixtures since it is necessary to capture individual weed spectra for separating Brazilian waterweed from

other submerged vegetation. In the future, it will be more helpful if points are collected for distinct patches of predominantly monospecific weeds.

The point sampling from within the patches was also ineffective in remote sensing identification of the SAV. The ReMetrix sampling protocol involved lifting the submerged vegetation from the channel bottom with the rake to the surface for identification and measurement. Plant samples collected from the bottom and within the water column are not representative of the vegetation at the water surface seen by the airborne instrument.

CONTINUED INVESTIGATIONS

Improving Current Delta-wide Classification

Our investigations will continue to improve the identification of target species within Priority 1 by creating a GIS network of the rivers and channels to separate tidal heights and water characteristics at the time of the overflights. Stratifying the Delta into channel and large water body segments of consistent water height and quality (sediment load, algae, etc.) should improve classification accuracy by reducing the variability in the water and vegetation spectra. These investigations will also allow us to better extend our mapping of the target species to areas not visited in field work, i.e., north of flightline number 23 and for the southern portion of Priority 1 below flightline number 55. Combining the water movement patterns with the GIS network may also be helpful in anticipating the spread of Brazilian waterweed under various natural and anthropomorphic environmental changes.

Mapping Invasive Species in Priority 2

Our work from January to June 2005, will focus on analysis of the imagery acquired for Priority 2. The techniques developed in Priority 1 will be improved and extended to the mapping the same SAV and water hyacinth. We will also investigate the potential for mapping pepperweed and purple loosestrife along the banks of the Tuolumne, Merced and San Joaquin Rivers. This effort will utilize the GPS locations of purple loosestrife provided by CDFG (Figure 26). While we acquired some field data for the Priority 2 region in early July, 2004, we do anticipate additional fieldwork to collect representative vegetation sites.

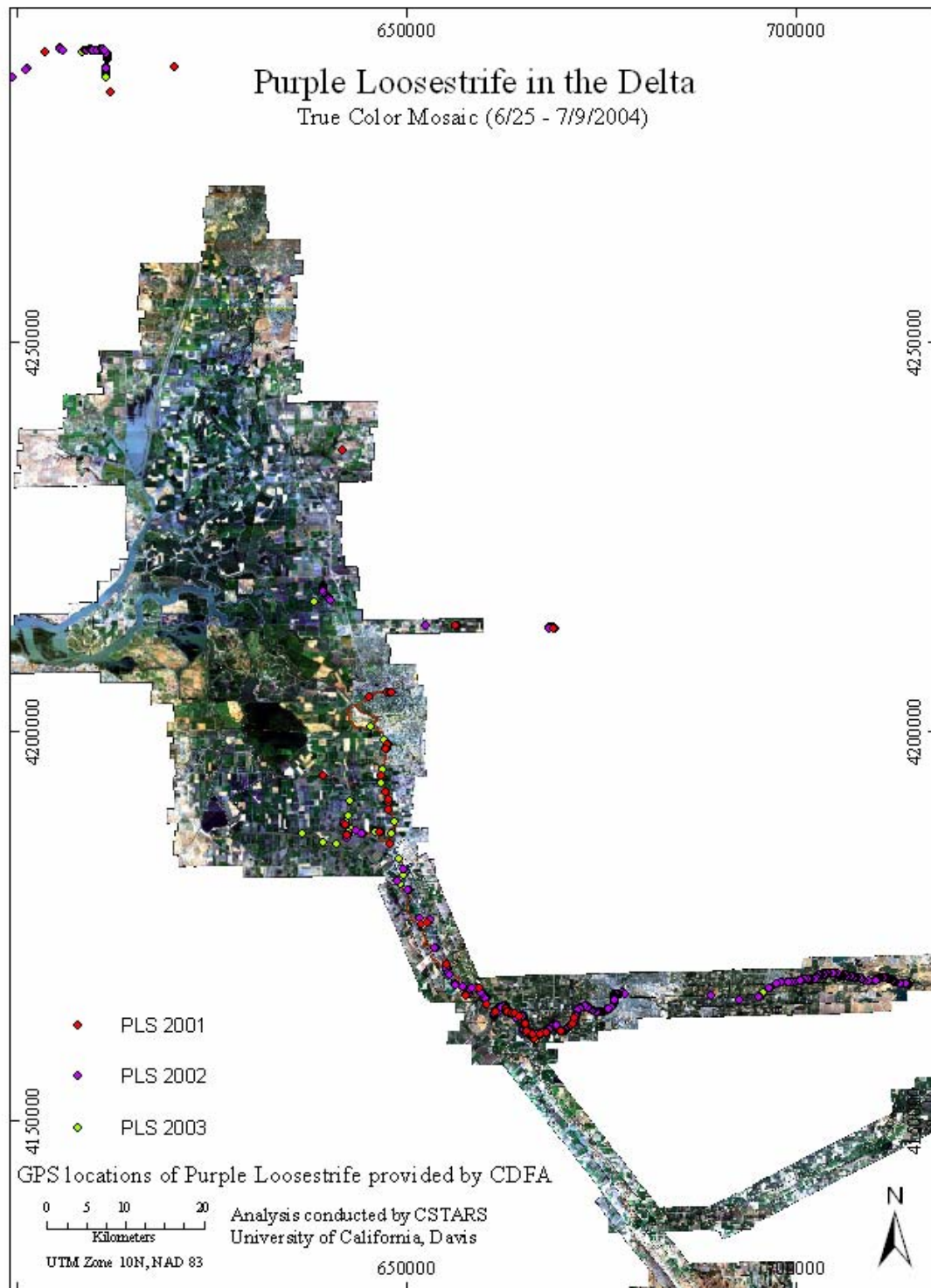


Figure 26

CONCLUSION

This study has expanded on the knowledge gained in the pilot project conducted in 2003 and extended the mapping of Brazilian waterweed and water hyacinth across 32 flightlines of the Delta. In addition, this work has classified these two target species and provided estimates of weed infestations at six priority sites identified by CDBW, and also made some preliminary investigations into the ability to distinguish Brazilian waterweed from other submerged aquatic species at one other site. The direct result of this research is an improved understanding of the distribution of Brazilian waterweed and water hyacinth throughout the Delta and provides a baseline dataset for the Department of Boating and Waterways and California Department of Food and Agriculture to monitor the spread of infestations as well as guidance for decision-making relating to control and eradication efforts within the Delta.

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REFERENCES

- Ackleson, S. G. and Klemas, V. (1987). Remote sensing of submerged aquatic vegetation in lower Chesapeake Bay: a comparison of Landsat MSS and TM Imagery. *Remote Sensing of Environment*, 22, 235-248.
- Adams, J. B., Sabot, D.E., Kapos, V., Almeida Filho, R., Roberts, D.A., Smith, M.O., and Gillespie, A.R. (1995). Classification of multispectral images based on fractions of endmembers: application to land cover change in the Brazilian Amazon. *Remote Sensing of Environment*, 52, 137-154.
- Adams, J. B., Smith, M. O., and Gillespie, A. R. (1993). Imaging spectroscopy: interpretation based on spectral mixture analysis. In: C. M. Pieters and P. A. J. Englert (Eds.), *Remote geochemical analysis: elemental and mineralogical composition* (pp. 145-156). Cambridge University Press, New York City, New York.
- Bossard, C. C., Randall, J. M., and Hoshovsky, M. C. (Eds.) (2000). *Invasive plants of California's wildlands*. University of California Press, Berkeley, California.
- Cocks, T., Jennsen, R., Stewart, A., Wilson, I. and Shields, T. (1998). *The HyMap airborne hyperspectral sensor: the system, calibration and performance*, Proceedings of the 1st EARSEL Workshop on Imaging Spectroscopy, Zurich, October 1998.
- Elmore, A. J., Mustard, J. F., Manning, S. J., and Lobell, D. B. (2000). Quantifying vegetation change in semiarid environments: precision and accuracy of spectral mixture analysis and the normalized difference vegetation index. *Remote Sensing of Environment*, 73, 87-102.
- Hortwitz, H. M., Lewis, J. T., and Pentland, A. P. (1975). Estimating proportions of objects from Multispectral Scanner Data. Final Report. NASA-CR 141862. 108pp.
- Johnson, D.E. (1999). Surveying, mapping, and monitoring noxious weeds on rangelands. In R. L. Sheley, and J. K. Petroff (Eds.), *Biology and management of noxious rangeland weeds* (pp. 19-35). Oregon State University Press, Corvallis, Oregon.
- Lehmann, A. and Lachavanne, J.B. (1997). Geographic information systems and remote sensing in aquatic botany. In: Lachavanne, J. B., Caloz, R., Lahmann, A. (Eds.), *Geographic information systems and remote sensing in aquatic botany. Aquatic Botany*, 58, 195-207.
- Lehmann, A., Jaquet, J. M., Lachavanne, J. B. (1997). A GIS approach of aquatic plant spatial heterogeneity in relation to sediment and depth gradients, Lake Geneva, Switzerland. In: Lachavanne, J. B., Caloz, R., Lehmann, A. (Eds.), *Geographic information systems and remote sensing in aquatic botany. Aquatic Botany*, 58, 347-361.
- Mooney, H. A. and Cleland, E. E. (2001). The evolutionary impact of invasive species. *Proceedures of National Academy of Science*, 98 (10), 5446-5451.

- Mulitsch, M. and S. Ustin, 2003. *Mapping Invasive Plant Species in the Sacramento-San Joaquin Delta Region Using Hyperspectral Imagery*. Report from the Center for Spatial Technologies and Remote Sensing at the University of California Davis, One Shield Avenue, Davis, California 95616. Report submitted to the California Department of Boating and Waterways, 2000 Evergreen Street, Sacramento, California 95815
- Penfound, W. and Earle, T. (1948). The biology of the water hyacinth. *Ecological Monographs*, 18: 447-72.
- Penuelas, J., Gamon, J. A., Griffin, K. L. and Field, C. B. (1993). Assessing community type, plant biomass, pigment composition, and photosynthetic efficiency of aquatic vegetation from spectral reflectance. *Remote Sensing of Environment*, 46, 110-118.
- Pimm, S. L., and Gilpin, M. E. (1989). Theoretical issues in conservation biology. In: Roughgarden, J., R. May and S.A. Levin (Eds.). *Perspectives in Ecological Theory*. Princeton University Press, Princeton, New Jersey. pp 287-205.
- Porter, D. E., Edwards, D., Scott, G., Jones, B., and Street, W.S. (1997). Assessing the impacts of anthropogenic and photographic influences on grass shrimp in localized salt-marsh estuaries. In: Lachavanne, J.B., Caloz, R., Lehmann, A. (Eds.), *Geographic Information Systems and Remote Sensing in Aquatic Botany*. *Aquatic Botany*, 58, 289–306.
- Riano, D., Chuvieco, E., Ustin, S., Zomer, R., Dennison, P., Roberts, D., and Salas, J. (2002). Assessment of vegetation regeneration after fire through multitemporal analysis of AVIRIS images in the Santa Monica Mountains. *Remote Sensing of Environment*, 79, 60-71.
- Roberts, D. A., Green, R. O., and Adams, J. B. (1997). Temporal and spatial patterns in vegetation and atmospheric properties from AVIRIS. *Remote Sensing of Environment*, 62, 223-240.
- Sanderson, E. W., Zhang, M, Ustin, S. L., and Rejmankova. E. (1998). Geostatistical scaling of canopy water content in a California salt marsh. *Landscape Ecology*, 13, 79-92.
- Sawaya, K. E., Olmanson, L. G., Heinert, N. J., Brezonik, P. L., and Bauer, M. E. (2003). Extending satellite remote sensing to local scales: land and water resource monitoring using high-resolution imagery. *Remote Sensing of Environment*, 88, 144-156.
- Schmidt, K. S., and Skidmore, A. K. (2003). Spectral discrimination of vegetation types in a coastal wetlands. *Remote Sensing of Environment*, 85, 92-108.
- Silvestri, S., Marani, M., Settle, J. Benvenuto, F., and Marani, A. (2002). Salt marsh vegetation radiometry: data analysis and scaling. *Remote Sensing of Environment*, 80, 473-482.
- Small, C. (2001). Estimation of urban vegetation abundance by spectral mixture analysis, *International Journal of Remote Sensing*, 22, 1305–1334.
- Small, C. (2002). Multitemporal analysis of urban reflectance, *Remote Sensing of Environment*, 81, 427– 442.
- Smith, M. O., Ustin, S. L., Adams, J. B., and Gillespie, A. R. (1990). Vegetation in deserts: I. A regional measure of abundance from multispectral images. *Remote Sensing of Environment*, 31, 1-26.

- Spanglet, H. J., Ustin, S. L., and Rejmankova, E. (1998). Spectral reflectance characteristics of California subalpine marsh plant communities. *Wetlands*, 18(3), 307-319.
- Ustin, S. L., Roberts, D. A., Jacquemoud S., Pinzon, J., Gardner, M., Scheer, G., Castaneda, C.M., and Palacios, A. (1998). Estimating canopy water content of chaparral shrubs using optical methods. *Remote Sensing of Environment*, 65, 280-291.
- Ustin, S. L., Hart, Q. J., Duan, L., and Scheer, G. (1996). Vegetation mapping on hardwood rangelands in California. *International Journal of Remote Sensing*, 17 (15), 3015-3036
- Valta-Hulkkone, D. Pellikka, P., Tanskanen, H., Ustinov, A., and Sandman, O. (2003). Digital false color aerial photographs for discrimination of aquatic macrophyte species. *Aquatic Botany*, 75, 71-88.
- Ward, D. H., Morton, A., Tibbitts, T. L., Douglas, D. C., and Carrera-Gonzalez, E. (2003). Long-term change in eelgrass distribution at Bahia San Quintin, Baja California, Mexico, using satellite imagery. *Estuaries*, 26, 1529-1539.
- Williams, A. P. and Hunt, Jr., E. R. (2002). Estimation of leafy spurge cover from hyperspectral imagery using mixture tuned matched filtering. *Remote Sensing of Environment*, 82, 445-456.
- Williams, D. C. and Lyon, J.G. (1997). Historical aerial photographs and a geographic information system (GIS) to determine effects of long-term water level fluctuations on wetland along the St. Mary's River, Michigan, USA. In: Lachavanne, J.B., Caloz, R., Lehmann, A. (Eds.), *Geographic Information Systems and Remote Sensing in Aquatic Botany*. *Aquatic Botany*, 58, 363-378.
- Zhang, M., Ustin, S. L., Rejmankova, E., and Sanderson, E.W. (1997). Remote sensing of salt marshes: potential for monitoring. *Ecological Applications*, 7(3), 1039-1053.
- Zhang, X. (1998). On the estimation of biomass of submerged vegetation using Landsat thematic mapper (TM) imagery: a case study of the Honghu Lake, PR China. *International Journal of Remote Sensing*, 19 (1), 11-20.

Appendix

Table 1. Final number of GPS points collected in fieldwork after screening (Priority 1 only). ‘SAV’ = Submerged Aquatic Vegetation, ‘EmergEd’ = EmergEd species, ‘Levee’ = Levee species.

Vegetation Lifeform	Species Common Name	Species Scientific Name	Number of GPS points
Water	None		86
SAV	Brazilian waterweed	<i>Egeria densa</i>	432
SAV	Algae		81
SAV	Watermilfoil	<i>Myriophyllum spicatum</i>	69
SAV	Coontail	<i>Ceratophyllum demersum</i>	29
SAV	Cabomba	<i>Cabomba caroliniana</i>	12
SAV	American pondweed	<i>Potamogeton nodosus</i>	25
SAV	Curly leaf pondweed	<i>Potamogeton crispus</i>	25
SAV	Pondweed spp	<i>Potamogeton</i> spp.	23
SAV	Common waterweed	<i>Elodea canadensis</i>	5
SAV	Mixed SAV		39
EmergEd	Water hyacinth	<i>Eichhornia crassipes</i>	262
EmergEd	Pennywort	<i>Hydrocotyle ranunculoides</i>	201
EmergEd	Water primrose	<i>Ludwigia peploides</i>	71
EmergEd	Azolla	<i>Azolla</i> spp.	12
Levee	Tule	<i>Scirpus acutus</i>	259
Levee	Cattail	<i>Typha latifolia</i>	54
Levee	Blackberry	<i>Rubus armeniacus</i>	31
Levee	Arundo	<i>Arundo donax</i>	33
Levee	Pepperweed	<i>Lepidium latifolium</i>	12
Levee	Common reed	<i>Phragmites australis</i>	26
Levee	Willow	<i>Salix</i> spp.	42
Levee	Live oak	<i>Quercus wislizeni</i>	12
Levee	Cottonwood	<i>Populus</i> spp.	7
Levee	Alder	<i>Alnus</i> spp.	1
Levee	Mixed grasses		34
Levee	Mixed riparian		2

Table 2. RMS error associated with Hymap flightline geocorrection to 1-foot color photos.

Flightline number	Overall RMS Error	Flightline number	Overall RMS Error
1	0.95	33	1.05
2	0.78	34	0.98
3	0.78	35	0.94
4	0.94	36	0.89
5	0.75	37	0.83
6	0.96	38	0.93
7	0.98	39	0.86
8	0.98	40	0.94
9	1.01	41	1.15
10	0.99	42	1.18
11	0.64	43	0.86
12	0.95	44	1.37
13	0.77	45	0.95
14	0.73	46	0.94
15	0.85	47	0.84
16	0.88	48	0.95
17	0.71	49	0.82
18	0.77	50	0.84
19	0.82	51	1.21
20	0.90	52	0.95
21	0.79	53	0.69
22	0.99	54	0.66
23	0.09	55	0.85
24	0.85	56	0.86
25	0.86	57	0.82
26	0.78	58	0.85
27	0.82	59	0.91
28	0.74	60	0.74
29	0.97	61	0.94
30	0.81	62	0.90
31	0.80	63	0.90
32	1.06	64	0.93